

Kelly Special Restoration Advisory Board (RAB) Technical Review Subcommittee (TRS) SUMMARY

November 9, 2004 Kennedy High School Cafeteria 1922 S. General McMullen Dr. San Antonio, TX 78226

1. Attendees:

Mr. Don Barker Ms. Sandra Converse Ms. Kyle Cunningham Ms. Larisa Dawkins Ms. Leigh-Ann Fabianke Dr. David A. Fowler Mr. Ramon A. Garcia Mr. Rodrigo Garcia Ms. Sarah Garcia Mr. Henry Galindo Ms. Esmeralda Galvan Ms. LeAnn Herren Mr. Brian Howard Ms. Jill Johnston Mr. Brian M. Kaplan Ms. Cheri Kirkpatrick Ms. Norma Landez Ms. Maria Teran-McIver Mr. Carlos San Miguel Mr. Gary Miller

Mr. Sam Murrah Mr. Pete Muzquiz Ms. Jannie O'Neal Mr. Nazarite Perez Ms. Abbi Power Mr. Armando Ouintanilla Mr. Genaro Rendon Mr. Sam Sanchez Mr. Jeffrey Shire Mr. Michael Sheneman Mr. Robert Silvas Mr. Kelley Siwecki Mr. Brendan Smith Dr. David Smith Mr. Tim Sueltenfuss Ms. Robyn Thompson Mr. Glenn Wilkenson Mr. James Wittmer Mr. David Yantz

- 2. Introduction. Dr. David Smith, RAB/TRS Facilitator, opened the meeting at 6:35 p.m. and called the roll. A quorum was achieved just after the meeting began. Dr. David Smith announced that Mr. William Ryan would be filling Mr. Adam Antwine's seat as government co-chair for this Special RAB meeting. The Pledge of Allegiance was said and then a moment of silence was observed. Dr. David Smith announced that the goal of the meeting was to review and comment on the Agency for Toxic Substances and Disease Registry (ATSDR) Health Consultation, Past Air Emissions Report for Kelly Air Force Base.
- **3.** ATSDR Briefing. Dr. David A. Fowler and Mr. Brian M. Kaplan of the ATSDR presented a briefing on the ATSDR Health Consultation, Past Air Emissions Report for Kelly Air Force Base.
- 4. Community Comment/Question-and-Answer Session on the ATSDR Past Air Emissions Report. Dr. David A. Fowler, Mr. Brian M. Kaplan and Ms. Maria Teran-McIver responded

to questions from the audience and RAB members in attendance concerning the ATSDR Report and related subjects.

- 5. Technical Assistance for Public Participation (TAPP) Update. Ms. Larisa Dawkins presented a briefing on what the TAPP is and how it works. She also provided a status update of the TAPP program budget.
- 6. Community Comment/Question-and-Answer Session on the TAPP update. Ms. Larisa Dawkins responded to questions from the audience and RAB members in attendance concerning the TAPP Update and related subjects.
- 7. Meeting Wrap-Up. Dr. David Smith stated that action items will be discussed at the next regularly scheduled RAB meeting. Meeting summaries will also be approved at the next regular RAB meeting. Dr. David Smith also reminded the RAB members and the community that RAB elections will take place at the January 18, 2005 RAB meeting.
- 8. Next Meeting. The next regularly scheduled RAB meeting is set for Tuesday, January 18, 2005, at 6:30 p.m. at Kennedy High School. The next regularly scheduled TRS meeting is set for Monday, December 13, 2004, at 6:30 p.m. at the Environmental Health and Wellness Center.
- 9. Adjourn. The meeting adjourned at 9:45 p.m.

November 2, 2004

Mr. Robert Silvas

Agency for Toxic Substances and Disease Registry Attention: Dr. David Fowler 1600 Clifton Road NE Mailstop E32 Atlanta, GA 30333

Re: ATSDR Past Air Emissions Study

Dr. Fowler,

On behalf of the Kelly Restoration Advisory Board (RAB), I am writing to express concern over the release of the past air emissions study information to the news media. Information on the study was published in two newspapers before it was released to the RAB. This puts pressure on the RAB to rush decisions to approve studies that have not been properly reviewed by the RAB. In the future, please wait until the RAB has had the opportunity to review the studies before releasing the information to the news media.

If you have any questions or would like to discuss this further, please contact me at (210) 340-0980. I look forward to your presentation at our next meeting on Tuesday, November 9, 2004, at 6:30 p.m. in the Kennedy High School cafeteria.

Sincerely, libor

Robert Silvas Interim RAB Community Co-chair

CC: RAB members and alternates Southwest Workers' Union U.S. Congressman Ciro D. Rodriguez U.S. Congressman Charles Gonzalez San Antonio City Councilwoman Patti Radle San Antonio City Councilwoman Richard Perez Mayor Ed Garza Alamo Area Council of Governments

Consejo Consultivo para la Restauración Especial de Kelly (RAB, por sus siglas en inglés) Subcomité de Revisión Técnica (TRS, por sus siglas en inglés) RESUMEN

9 de noviembre de 2004 Cafetería de la Preparatoria Kennedy 1922 S. General McMullen Dr. San Antonio, TX 78226

1. Asistentes:

Sr. Don Barker Sra. Sandra Converse Sra. Kyle Cunningham Sra. Larisa Dawkins Sra. Leigh-Ann Fabianke Dr. David A. Fowler Sr. Ramon A. Garcia Sr. Rodrigo Garcia Sra. Sarah Garcia Sr. Henry Galindo Sra. Esmeralda Galvan Sra. LeAnn Herren Sr. Brian Howard Sra. Jill Johnston Sr. Brian M. Kaplan Sra. Cheri Kirkpatrick Sra. Norma Landez Sra. Maria Teran-McIver Sr. Carlos San Miguel Sr. Gary Miller

- Sr. Sam Murrah Sr. Pete Muzquiz Sra. Jannie O'Neal Sr. Nazarite Perez Sra. Abbi Power Sr. Armando Quintanilla Sr. Genaro Rendon Sr. Sam Sanchez Sr. Jeffrey Shire Sr. Michael Sheneman Sr. Robert Silvas Sr. Kelley Siwecki Sr. Brendan Smith Dr. David Smith Sr. Tim Sueltenfuss Sra. Robyn Thompson Sr. Glenn Wilkenson Sr. James Wittmer Sr. David Yantz
- 2. Introducción. Dr. David Smith, Intermediario de RAB/TRS, hizo la apertura de la junta a las 6:35 p.m. y pasó lista. Se alcanzó el quórum exactamente después de que empezó la junta. El Dr.David Smith anunció que el Sr. William Ryan ocuparía el puesto del Sr.Adam Antwine como co-dirigente del gobierno para esta junta Especial de RAB. Se hizo el Juramento de Lealtad y después se guardó un momento de silencio. El Dr. David Smith anunció que la meta de la junta era revisar y comentar sobre la Consulta de Salud de la Agencia para Sustancias Tóxicas y Registro de Enfermedades (ATSDR, por sus siglas en inglés), el Reporte Pasado de Emisiones de Aire de la Base Kelly de la Fuerza Aérea.
- **3. Informe del ATSDR.** El Dr. David A. Fowler y el Sr. Brian M. Kaplan de ATSDR presentaron un informe sobre la Consulta de Salud de ATSDR, el Reporte Pasado de Emisiones de Aire de la Base Kelly de la Fuerza Aérea.
- 4. Comentarios de la Comunidad / Sesión de Preguntas y Respuestas sobre el Reporte Pasado de Emisiones de Aire de ATSDR. El Dr. David A. Fowler, el Sr. Brian M. Kaplan y

la Sra. Maria Teran-McIver respondieron las preguntas de la audiencia y de los miembros de RAB que asistieron, concernientes al Reporte de ATSDR y asuntos relacionados.

- **5.** Actualización de la Asistencia Técnica para la Participación Pública (TAPP, por sus siglas en inglés.) La Sra. Larisa Dawkins presentó un informe sobre lo que es TAPP y cómo funciona. También proporcionó una actualización de la situación del presupuesto del programa TAPP.
- 6. Comentarios de la Comunidad / Sesión de Preguntas y Respuestas sobre la actualización de TAPP. La Sra. Larisa Dawkins respondió las preguntas de la audiencia y de los miembros de RAB que asistieron, concernientes a la Actualización de TAPP y asuntos relacionados.
- 7. Terminación de la Junta. El Dr. David Smith estableció que las acciones a tomar, serán discutidas en la siguiente junta de RAB programada regularmente. Los resúmenes de la junta también serán aprobados en la siguiente junta regular de RAB. El Dr. David Smith también les recordó a los miembros de RAB y a la comunidad, que las elecciones de RAB se llevarán a cabo en la junta de RAB del 18 de enero de 2005.
- 8. Próxima Junta. La siguiente junta de RAB programada regularmente, será el martes 18 de enero de 2005, a las 6:30 p.m. en la Preparatoria Kennedy. La siguiente junta de TRS programada regularmente será el lunes 13 de diciembre de 2004, a las 6:30 p.m. en el Centro de Higiene y Bienestar Ambiental.
- 9. Cierre. Se levantó la sesión a las 9:45 p.m.

Agency for Toxic Substances and Disease Registry

Kelly Air Force Base Past Air Emissions

> David A. Fowler, PhD Brian M. Kaplan, MS. MA

> > ATSDR=1

Why estimate past air emissions?

- Original health assessment identified elevated liver cancer, birth defects, low birth weights, and leukemia
- Due to latency, potential environmental exposures would have occurred years prior to clinical disease

ATSUR-2





- Charles Williams, AFBCA-Kelly
- Everett Douglas, Navy AESO
- Vincent Tino/George Siple, CDM Federal Programs

ATSOR

Reconstruct Past Air Concentrations (Air Dispersion Modeling)

- Emission locations (industrial and aircraft)
- Emission parameters (chemical and physical)
- Model (ISCST3) and model parameters

ATSDR

Uncertainties and Limitations

- We used a model (mathematical) approximations)
 - ISCST3 expect ±100% (i.e. 1 μg/m³ likelÿ represents 0.5 to 2 μg/m³)
- Model inputs (emissions data)
- Emissions averaged over a year

ATSOR-

Information Sources

- Kelly Air Force Emissions
 - Stationary (Kelly emissions inventories) Past (1970 to 1975, 1983 to 1989); Current (1995)
 - Aircraft (Specific data and general studies)
 Aircraft; engine testing; mumber of operations
- Meteorological Data
 - Surface data-San Antonio International Airport
 - Upper air data-Del Rio International Airport

ATSUR



Stationary (Industrial) Sources

- Temperature of exit gas (defaults-68°F)*
- Diameter of stack at exit (3.2 ft)
- Exit gas velocity (0.33 ft/s)
- Release or stack height (20 ft)
- Rate of release
- Location

ATSDR





Aircraft Toxic Air Dispersion Modeling

- Types of planes/engines
- Number of landing-takeoff (LTO) and touch and go operations—frequency
- Locations of planes
- Emission factors/Time-in-modes

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Types of planes/engines

A-10,A-4,A-6,B-1,B-17,B-24,B-29,B-36,B-50,B-52,B-58, B-727,B-757,C-118,C-119,C-12,C-121,C-123,C-124,C-130, C-141,C-17,C-17,C-2,C-21,C-47,C-5,C-54,C-7,C-74,C-9, C-97,DC-9,F-100,F-101,F-102,F-104,F-105,F-106,F-14, F-15,F-16,F-18,F-4,F-80,F-84,F-86,F-89,F-94,F8-111, KC-10,KC-135,P-38,P-47,P-51,SR-71,SW-4,SWB,T-1, T-28,T-33,T-34,T-37,T-38,T-39,T-41,T-43,T-45,T-6, UH-1,XC-99.

Joint Use Supplemental Environmental Impact Statement, August 2000 and Kelly's Official Pictural Matery Book (Fittp://1982.14.23.23.01) (Estorybook Intrafacested 2/2/2001.3)

Âtsdr-



Aircraft Operations-Kelly AFB







4500 N	1570 fj
1.5 miles	5.7 miles
	Atsor

Emission Factors

Chemicals of concern in exhaust from the TF33-P3 (B5211) engine using JP-4 jet fuel

			Power :	Setting				
	Idle		30%		75%		100%	
		%wt	%wt		%wt		%wt	
	ppmC	HAP/HC	ppmC	HAP/HC	ppmC	HAP/HC	ppmC	HAP/HC**
1,3-Butadiene	11,981	0.00920	0 571	0 007 10	0 0 2 4	0 00387	<0.001	0.00030
Benzene	12499	0 00924	1 698	0 02032	0 16	0.02481	0.029	0.00849
Naphthalene	10 138	0 00738	0 395	0 00465	0 0 3 5	0 00534	0 05	0 01442
Formaldehyde	15 54	0 0 2650	4 009	0 11068	0.423	0,15126	0.083	0.05607
Acetaldehyde	1 802	0 00226	1 564	0.03168	0211	0 05535	0 036	0 01784
Acrolein	1 833	0 00195	0 501	0 00861	0.051	0.01135	<0 001	0 00042
*C W Spicer, M W Holdren, S.E. Miller, D.L. Smith, R.N. Smith, D.P. Hughes "Aircraft Emissions Characterization," Final Report.								

Mamh 1988, Engineering and Services Laboratory, Air Force Engineering & Services Center, Tyndall Air Force Base, ESL-TR-87-63 **Non-detects were converted to % weight based on the detection level of 0 001

ATSOR-19

Time-in-mode

Operating pursmeters for ball and a Trut- 3 cagine.

Aircraft Mode	Engine	Minutes*	Fuel FlowPer Engine (1000 lb / hr)**	HC Emissions Per Engine (lb/1000 lb fuel)**
Startup	ldly	20	1.052	94.00
Outbound Taxi	Idle	9	1 052	94 00
Engine Check	Military	4.5	7 105	0 03
Runway roll	A/B*	07	7 105	0.03
Climbout I	A/B"	07	7,105	0 03
Climbout II	Military	08	7 105	0.03
Approach I	Idle	3	1 052	94.00
Approach II	Idle	1	1 0 5 2	94.00
Landing on runway	1die	1	1.052	94.00
Inbound Taxi	Idle	12	1.052	94_00
Idle at shutdown	Idle	48	1.052	94_00

*USAF Aircraft Pollution Emission Factors and Landing and Takeoff (LTO), Dennis Naugle, et al,

AD/A-00C 239 (February 1975) ****Aircraft Emissions Characterization," <u>C.W. Spicer, M.W. Holdren, S.E.</u> Miller, D.L. Smith, R.N. Smith, D.P. Hughes, Final Report, March 1988, Engineering and Services Laboratory, Air Force Engineering & Services Center, Tyndall Air Force Base, ESL-TR-87-63

ATSDR







ATSDR- 2







Findings

- Exposure to individual chemicals not likely to result in health effects
- Effects from exposure to the *mixture* of chemicals is indeterminate.
- Not enough information on hex chromium (before 1980), misting, or incineration

Why "indeterminate"?

Because of what we do not know about

- Data
- Exposure
- Toxicology/Epidemiology

ATSDR.

ATSOR 2

Next step

- If very high levels had been estimated, what would we do?
- Look to see if there were the expected health effects in the exposed population.

Key recommendation

 Investigate, the elevated leukemia outcomes

> ATSDR is currently investigating and will discuss in next meeting

ATSOR -

ATSOR-30

Next meeting

- Recap events since petition
- Update on East Kelly health assessment
- Update on health outcome data

ATSDR-31

ATSDR - Kelly Air Force Base, San Antonio, Texas



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ATSDR MEDIA ANNOUNCEMENT

ATSDR Releases for Public Review and Comment the Kelly AFB Public Health Consultation About Past Air Emissions

Public Comment Deadline is Nov. 30, 2004

For Immediate Release: October 22, 2004

ATLANTA - The public comment period for the just released Kelly Air Force Base (AFB) public health consultation about off-base air emissions runs through Nov. 30, 2004. The report was issued by the Agency for Toxic Substances and Disease Registry (ATSDR), a federal public health agency.

The health consultation is the eighth ATSDR investigation into possible exposure and health effects from environmental contamination related to the former Air Force base. The health consultation has four conclusions:

— Air dispersion modeling indicates that aircraft emissions of JP-4 jet fuel were unlikely to have resulted in off-base exposures to individual chemicals at levels that would cause harmful health effects.

— Insufficient data were available to determine the health hazard of exposure to hexavalent chromium air emissions. Off-base exposures to estimated individual contaminant levels of other chemicals emitted from stationary sources are unlikely to have caused adverse health effects.

— Off-base exposure to chemical mixtures from stationary and aircraft sources is an indeterminate health hazard because of the scientific uncertainty of potential interactions from exposure to chemical mixtures.

— Data were not available for ATSDR to evaluate potential exposure to emissions from incineration of cyanide wastes or to unburned, airborne aircraft fuel emissions (misting).

The public health consultation is available for review in San Antonio at

John F. Kennedy High School Library 1922 S. General McMullen

Pan American Library 1122 W. Pyron

Las Palmas Library 515 Castroville Road

Memorial Library 3222 Culebra The public comment period extends through Nov. 30, 2004. Comments on the public health consultation must be made in writing. Mail comments to Chief, Information Services Branch ATSDR 1600 Clifton Road, N.E. (MS E-60) Atlanta, GA 30333

Comments received during the public comment period will be logged in to the ATSDR administrative record for this health consultation. Comments received, without the names of individuals who submitted them, and ATSDR responses to the comments will appear in an appendix to the final public health consultation. Names of those who submit comments, however, will be subject to release for requests made under the U.S. Freedom of Information Act.

For more information, community members can contact Environmental Health Scientist Susan Moore or Health Communication Specialist Maria Teran-MacIver, toll free, at 1-888-422-8737. Senior Regional Representative George Pettigrew also may be contacted at 214-665-8361. Callers should refer to the Kelly Air Force Base site in San Antonio, Texas.

ATSDR, a federal public health agency of the U.S. Department of Health and Human Services, evaluates the human health effects of exposure to hazardous substances.

Established by Congress in 1980 under the Superfund law, ATSDR conducts public health assessments at each of the sites on the EPA National Priorities List, as well as other sites when petitioned.

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Members of the news media can request an interview with ATSDR staff by calling the ATSDR Office of Communication at 404-498-0070.

Back to ATSDR Home page

Updated by R. Searfoss October 28, 2004 For more information, contact ATSDR at: 1-888-422-8737 or <u>e-mail</u> (public inquiries) 404-498-0080 or <u>e-mail</u> (news media)

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http://www.atsdr.cdc.gov/NEWS/kellyairforce102204.html

Health Consultation Past Air Emissions Kelly Air Force Base San Antonio, Bexar County, Texas CERCLIS Number TX2571724333

Prepared by

Agency for Toxic Substances and Disease Registry Division of Health Assessment and Consultation Exposure Investigations and Consultations Branch

Forward

ATSDR was requested to evaluate potential exposure to past air emissions. To fulfill this request, ATSDR gathered relevant data and information to reconstruct potential past inhalation exposures. The available information from industrial activities and aircraft emissions was often scarce or non-existent. To provide the community and health officials with some perspective on potential past exposures, ATSDR performed a dose reconstruction by modeling the available information. ATSDR recognizes the estimates provided have a varying degree of uncertainty and caution should be exercised in the application of the estimates. This document describes the information used to estimate the exposures to past air emissions from Kelly Air Force Base. Recommendations are also included that provide public health follow-up activities that ATSDR considers prudent based on the results of the modeling effort and ATSDR's public health evaluation.

Information in this document is organized to improve readability by the public by placing methodology and scientific details in appendixes. The main body of the document contains the summary of the public health evaluation with supporting information contained in the appendixes.

Summary

ATSDR completed Phase I of the public health assessment (PHA) of Kelly Air Force Base (AFB) in August 1999 [1]. In Phase I, ATSDR recommended further investigation of potential exposures to past air emissions to be performed during Phase II. This health consultation is a part of Phase II and reports the evaluation of potential past exposures to air emissions from activities at Kelly AFB (see Table 1). This report was revised in January 2004 based on external peer review comments (see Appendix D for comments and responses).

Findings:

Off-base exposures to estimates of <u>individual contaminant levels</u> of hazardous air pollutants (HAPs) from **stationary** source emissions were unlikely to have resulted in adverse health effects and present **no apparent health hazard**. Data from past hexavalent chromium air emissions (before 1980) were insufficient to assess public health implications and represent an **indeterminate health hazard**.

Off-base exposures to estimates of <u>individual contaminant levels</u> from **aircraft** emissions were unlikely to have resulted in adverse health effects and present **no** apparent health hazard.

The uncertainty in potential interactions from off-base exposure to chemical mixtures from stationary and aircraft emissions represents an **indeterminate** *health hazard*.

Data were unavailable to evaluate potential exposure to emissions from incineration of cyanide wastes and fuel emissions from misting.

These findings are based in part on emissions inventory data, estimated air concentrations from air dispersion modeling, and toxicological data. The uncertainties of these data are discussed in this report and considered in these findings.

Exposure Pathway Elements								
Pathway	Contaminants	Source	Environmental Media	Point of Exposure	Route of Exposure	Exposed Population	Time	Comments
Air	Benzene 1,3-Butadiene Formaldehyde	Stationary Sources and Aircraft Emissions	Ambient Air	Off Base	Inhalation	Adult/Child	Past (before 1995)	Indeterminate health hazard to <u>cumulative</u> exposures of chemical <u>mixtures</u>
Air	Hexavalent Chromium	Stationary Sources	Ambient Air	Off Base	Inhalation	Adult/Child	Past ^a (before 1980)	Indeterminate
Air	Hexavalent Chromium	Stationary Sources	Ambient Air	Off Base	Inhalation	Adult/Child	Past (1980 and later)	No Apparent Health Hazard
Air	Individual HAPs ^b	Stationary Sources	Ambient Air	Off Base	Inhalation	Adult/Child	Past (before 1995)	No Apparent Health Hazard
Air	Individual Contaminants in JP-4 Jet Fuel Exhaust	Aircraft Emissions	Ambient Air	Off Base	Inhalation	Adult/Child	Past (before 1995)	No Apparent Health Hazard.
Air	Fuels, HAPs	Stationary Sources and Aircraft Entissions	Ambient Air	On Base	Inhalation	Worker	Past	ATSDR does not evaluate worker exposures. Recommendations are made for investigation by others.

a. Hexavalent chromium was emitted from 5 plating shops. The most significant were located in Buildings 258/295 and Building 301. Buildings 258/259 began operation in 1942 and shutdown in 1977. Building 301 replaced Building 258/259 in 1977. The emission rates of hexavalent chromium from Building 258/259 are not known. The emission rates from Building 301 are based on stack tests completed in 1980. The time prior to the 1980 stack test is used to define past exposures because of the unknown emission rates from Building 301 prior to 1980 and unknown emission rates from Buildings 258/259.

b. Hazardous Air Pollutants - see text for discussion.

4

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Background

The late Congressman Frank Tejeda requested that ATSDR investigate the potential relationship between environmental contaminant releases from Kelly AFB and the adverse health effects reported by area residents north and southeast of the base [2]. ATSDR publicly released findings during Phase I on August 24, 1999, and also described activities to be performed during Phase II. During Phase I, ATSDR performed an air dispersion screening model of air emissions from stationary sources to estimate possible air contaminant concentrations in the community. ATSDR concluded there was no apparent public health hazard to the community from exposure to current air emissions (1995 and after). However, the available information was inadequate to evaluate the potential for health effects from exposures to *past* air emissions (before 1995).

ATSDR considers past air emissions to have been an important contributor to potential environmental contamination and past exposure because:

- pollution control measures were not closely regulated and pollution control may not have been routinely used,
- the use of toxic chemicals in the workplace was more prevalent because of the limited knowledge of environmental health effects, and
- JP-4 jet fuel was used until 1994 at Kelly AFB and contained an average benzene concentration greater than the benzene concentration of the currently used jet fuel, JP-8 [3].

The following issues related to past air emissions are addressed in this document:

Stationary source emissions

- stationary emissions from processes such as chromium plating, painting, and degreasing.
- incinerator emissions involving cyanide (requested by the community).

Aircraft activity emissions (from mobile sources)

- emissions during takeoff, landing, and taxi operations.
- the reported "misting" through the inefficient burning of jet fuels during and previous to the 1970s (reported by Kelly Air Force Base [4]).

Results and Discussion

How does ATSDR evaluate past air emissions?

- Available air emissions data are evaluated to determine the public health implications from potential exposures. The uncertainty in available data, air sampling, air dispersion modeling methodology, and exposure information may vary. These uncertainties result in a varying degree of confidence in the conclusions.
- Air quality sampling and analysis have only been generally available for the more recent past. For distant past air emissions, air dispersion modeling is an important tool available to estimate concentrations that may have been present in the community. Air dispersion modeling can estimate the location and concentration of air contaminants released by the source of interest. Air dispersion modeling can also differentiate emissions from the source of interest and emissions from other sources, such as other industrial sources and automobiles. Air dispersion modeling, as with ambient air monitoring, do not determine a persons exposure because of a person's movement throughout the day. Air concentrations from air dispersion modeling are considered estimates because they are calculated values using mathematical formulas representing the atmosphere. These calculations introduce some uncertainty which are considered in the evaluation. The uncertainty in the location of a predicted concentration is often higher then the actual value (i.e., models are good for determining the air concentrations, but not exact locations). Because of the lack meteorological and emissions detail, models are not good at determining short term episodic events.
 - All available information is used to make conclusions about site-specific exposures. The estimated contaminant levels are compared to health-based comparison values derived by ATSDR, the Environmental Protection Agency, or state environmental and health agencies. Exposure to these levels would not be expected to result in adverse health effects, even for sensitive people in the general population. If an individual contaminant level does not exceed health-based comparison values, no further analysis of exposure to that individual contaminant is needed; however, the contaminants may be included when considering chemical mixtures or cumulative analysis. If a contaminant exceeds health-based comparison values, ATSDR performs further analysis including a risk analysis. Risk analysis is a multidimensional endeavor and may include a risk assessment, a toxicological evaluation, and an evaluation of health outcome data and epidemiological studies. Professional judgment is used to reach conclusions and make recommendations which may include follow-up activities such as health education, health studies, and public health interventions [5, 6].

How did ATSDR evaluate past air emissions at Kelly AFB?

ATSDR requested data from Kelly AFB for representative past air emissions, but routine sampling and analysis data of ambient air emissions were not available for the past era of concern. An air dispersion model of these emissions data estimated contaminant levels that may have been present in the community. Contaminants were selected for investigation by considering both toxicity and quantity used or emitted. Contaminants whose past use was similar to current use were not modeled during Phase II if the Phase I modeling did not suggest a public health concern and the source location or stack height were also similar. Appendix A addresses the air dispersion modeling methodology.

The modeling of aircraft emissions is a complex task. Many different aircraft must be considered as well as the numbers and types of engines used on specific aircraft. The engine efficiency, burn temperatures, and operating modes may be different for different types of engines, resulting in different emissions during operations such as taxi, take off, afterburn, approach, and landing. Emission estimates are further compounded by different flightline use in different years by different aircraft. Data were not available to complete the input values needed for this complexity of modeling. Therefore, ATSDR performed an air dispersion model on a worst-case scenario to estimate whether these emissions could be of public health concern. For the worst-case scenario, ATSDR selected these modeling inputs:

- the aircraft having the most engines,
- the least efficient engine for modeling of emissions,
- the engine with the highest emissions, and
- a year in which operations were the highest reported.

ATSDR also performed a dispersion model for a scenario representing planes with lower emissions. The modeling assumptions and specific model input parameters are provided in Appendix B.

ATSDR performed an air dispersion model using U.S. EPA's Industrial Source Complex Short Term Version 3 model (ISCST3) for both stationary source (industrial) and aircraft emissions. Both stationary and aircraft emissions were used to estimate individual contaminant levels and subsequent risk in the community. Contaminants with the highest estimated chronic risk (considering quantity and toxicity) were selected for evaluating chemical mixtures and cumulative exposures.

The ISCST3 model in flat terrain similar to Kelly Air Force Base has been shown to be accurate within two-times to one-half the actual result [7]. For instance, if the "real" value is $1 \mu g/m^3$, the model could show a range of $2 \mu g/m^3$ to $0.5 \mu g/m^3$. The largest uncertainty in this study, though, is the emission data used in the model and which are not accounted for in this error range.

8

What did ATSDR find?

Data Acquisition

Information about stationary (industrial) emissions and incinerator emissions involving cyanide wastes was requested from Kelly AFB. In addition, information about issues related to aircraft emissions, including speciated aircraft emissions using JP-4 jet fuel and aircraft misting (as described by Kelly AFB) was requested.

In March 2000, Kelly AFB submitted a report containing data and information about stationary and aircraft emissions [8]. Clarification and explanation of these data and information was requested. Kelly AFB submitted additional explanation in June 2000 [9]. ATSDR requested further clarification and explanation of both the original (March 2000) and updated data and information (June 2000), which Kelly AFB submitted in December 2000 [10]. Kelly AFB reported that some of the data and information requested could not be located. The available data are not comprehensive and may not be representative of past air emissions.

Sufficient data were acquired for:

stationary emissions (except for hexavalent chromium) aircraft emissions

Sufficient data were not acquired for:

past air emissions of hexavalent chromium air emissions due to "misting" incinerator emissions involving cyanide wastes

Stationary Emissions

Industrial Sources

For industrial activities except chromium plating and cyanide incineration, the data supplied by Kelly AFB were sufficient for analysis and making conclusions. Data were provided for the following contaminants: tetrachloroethylene (PCE), hexavalent chromium, methylene chloride, methyl ethyl ketone, benzene, ethyl benzene, formaldehyde, toluene, xylene, styrene, naphthalene, acrolein, acetaldehyde, trichloroethylene (TCE), trichloroethane (TCA), and dichloroethane (see Appendix B, Attachment 1, for a listing of chemicals modeled, locations, and emission rates) [8]. ATSDR performed an air dispersion model of these emissions and found that the *annual average <u>maximum</u> off-base concentrations of most chemicals did not exceed health-based comparison values. No chemicals exceeded noncancer comparison values. The maximum off-base concentrations of two chemicals (PCE and methylene chloride) exceeded a cancer comparison value and required further analysis (Appendix B, Table B-1). Hexavalent chromium data from plating operations were insufficient for evaluation. See Appendix B for more detail.*

Using modeling and analysis, ATSDR concluded that estimated levels of individual contaminants in the community would not represent a public health hazard. However, insufficient data were provided for evaluation of hexavalent chromium.

Incineration of cyanide waste

Kelly AFB reported that the incinerator that burned cyanide waste operated for about a year, but never operated properly [11, 12]. Kelly AFB did not submit quantitative data regarding the incineration of cyanide waste. Therefore, insufficient information is available for a health evaluation of potential exposure to cyanide air emissions from incineration.

Aircraft Emissions

Speciated jet fuel emissions.

ATSDR requested speciated JP-4 jet fuel emissions data and aircraft operational information such as takeoffs, landings, and taxi activities. JP-4 jet fuel was used until 1994 when the base converted to JP-8 jet fuel [13]. JP-4 jet fuel may have contained 100 times more benzene than JP-8 jet fuel [3] Kelly AFB provided information on the speciation of emissions of JP-8 jet fuel and on volatile organic chemicals (VOCs), nitrogen-oxygen compounds (NOx), and sulfur-oxygen compounds (SOx). The speciation of fuel was important as speciation identifies the individual chemicals present, such as benzene, enabling ATSDR to perform evaluations on specific chemicals. The information on speciation of emissions from aircraft using JP-4 jet fuel acquired by ATSDR was difficult to find and may not be representative of specific aircraft emissions from Kelly AFB activities. Current and past operational data were provided by Kelly AFB and consisted of numbers of takeoffs and landings [14, 15]. Data on JP-4 jet fuel speciation acquired by ATSDR and operational data provided by Kelly AFB were used to conduct an air dispersion model of aircraft emissions. A worst-case jet fuel emissions scenario was used for modeling aircraft emissions. The Industrial Source Complex air dispersion model was used (ISCST3, see Appendix B for details).

The modeling scenario included 336,000 takeoff and landings per year of a B52 (which has eight engines) using the least efficient engine (TF33-3). This modeling effort identified a potential worst-case scenario that would overestimate emissions. To give some perspective of the conservative nature of this approach, ATSDR also modeled emissions from an F16 aircraft, which has only one engine. A B52 emits an estimated 16 times more 1,3-butadiene and 8 times more benzene than an F16 (with an F110 engine) during takeoff and landing operations [16]. The operational data used in the model were about 3 times the average operational data after 1973 (330,000 operations in 1964. See Appendix B, Figure B-1). Concentrations estimated by the air dispersion model were the annual averages of the maximum off-base concentrations. Benzene, 1,3-butadiene, and formaldehyde contributed the highest estimated risk (see Appendix B for discussion).

Estimated levels using a worst-case scenario indicated that past air emissions of <u>individual</u> contaminants from aircraft would <u>not</u> be cause for public health concern. See Appendix B for more detail.

Misting

During the Viet Nam years, area residents described frequently experiencing a mist of jet fuel which they attributed to fuel jettisoning. ATSDR evaluated fuel jettisoning during the Phase I

Public Health Assessment and found that the Air Force did not keep records of fuel jettisoning. A frequent experience of mist in the community would be unlikely if Air Force policy concerning fuel jettisoning were followed. Kelly AFB identified another potential cause for the jet fuel mist experienced by the community in comments to the Phase I PHA, as follows:

"The flight aircraft's of the 50's, 60's, and early 70's routinely sprayed minor droplets of unburned fuel on approach and departure ends of the runways. The engines were not as efficient as today's engines. The amount of the spray was small, but could have been noticed as a very fine mist. Further, C-5s were not actually deployed until 1973. Most unburned fuel evaporated shortly after being blown out of the tailpipe. This spray was usually attributed to aircraft using afterburns, as after-burners function by dropping large amounts of fuel in the burn basket. A minuscule amount of fuel does not burn completely" [4].

Kelly AFB personnel were unable to locate quantitative information on misting. Kelly AFB has prepared a qualitative assessment to address this issue [17]. (See Appendix B, Attachment 2). ATSDR's investigation indicates that the mist that residents recall may not have been due to "misting", as defined by the Kelly AFB comment. NASA and USAF scientists report that they have not encountered the above phenomena. Exhaust temperatures are in excess of 400 degrees Celsius, one meter behind the exit plume of an F-14 [18]. All fuel should be in the gaseous form at this temperature. Other possible reasons for misting are speculated to be caused by leaking fuel, improperly jettisoned fuel, fuel jettisoned during an emergency, or condensation.

Takeoff and landings during the 1960s have been reported to be greater than 300,000 operations per year, which would be equivalent to one operation every two minutes on a 24-hour basis [14]. (An operation is assumed here as one takeoff or one landing, and one touch and go maneuver is counted as two operations. This assumption is discussed in Appendix B). At this high rate of activity, it is conceivable that the combination of fuel leakage, inefficient burning, and improperly jettisoned fuel from individual aircraft could have a cumulative effect on ambient air quality, especially if the majority of operations were performed during daylight hours. However, the lack of data precludes a quantitative evaluation by ATSDR.

Cumulative Assessment and Chemical Mixing

The limited data from past air emissions from Kelly AFB are not adequate to address comprehensive cumulative risks because adequate data are not available on all contaminants. Nevertheless, ATSDR performed an assessment based on the available data and current scientific literature. Where appropriate, ATSDR assumed that available data were representative of past air emissions. The uncertainty in the limited available data, the air dispersion model, estimates of potential exposures, and the cumulative effects of chemical mixtures suggests little confidence in the comprehensiveness of such an effort.

Individuals come into contact with chemicals identified at Kelly AFB and other chemicals through non-site-related exposures in the environment, home, and workplace. An individual may be exposed to chemicals in many ways including in medicines, food, vehicle exhaust, alcohol, and drinking water. The total exposure that an individual experiences, as well as individual risk factors, determine if a person has health effects resulting from the exposures. The best cumulative risk assessment given today's state of the science would fall short of being able to include an evaluation of the magnitude and interactions of *all* stressors and effects. At best, the risk estimates of a cumulative risk assessment will reflect *some* of the risks which may be reflected in community health statistics.

Exposure estimates of cumulative risks from aircraft and industrial emissions suggest a *moderate* cumulative risk for developing cancer <u>if animal data are used</u> and a *low* cumulative risk for developing cancer <u>if human data are used</u> exclusively (see Appendix C, Table C-1). ATSDR assigns a higher weight to well-designed and well-executed epidemiologic (human) studies than to animal studies of comparable quality in evaluating the potential human cancer risks. Epidemiological studies of occupational exposures suggest that exposures to 1,3-butadiene and benzene at air concentrations much higher than those estimated around Kelly AFB may be associated with the development of leukemia [19, 20]. However, workers are considered the healthiest segment of the general population. The levels at which other segments of the population might be effected is unknown. In addition, these occupational studies reported numbers of leukemia mortality (death) and not numbers of people developing the disease (incidence) or adverse health effects.

Formaldehyde has been associated with leukemia mortality in embalmers but not in industrial environments [21–24]. The differences in metabolism and mode of action does not suggest that formaldehyde would contribute to potential cumulative effects from exposures to benzene and 1,3-butadiene.

Through air dispersion modeling, ATSDR identified the community areas where exposure to the highest concentrations was most likely. Although not comprehensive, ATSDR can evaluate biologically plausible health outcome data to determine whether these health outcomes are occurring in this population at rates similar to or different from the general population. For some health outcomes, ATSDR further evaluated whether an association with an environmental exposure to air emissions from Kelly AFB was plausible (See ATSDR Health Outcome Data Evaluation Health Consultation [25]).

ATSDR investigated biologically plausible health outcome data in the 1999 Public Health Assessment [1]. Results of the investigation revealed some elevated health outcomes were not likely to be associated with an exposure to known contaminants from Kelly AFB. However, some elevated health outcomes could not be ruled out as having been associated with contaminants from Kelly AFB.

ATSDR concluded that some plausible cancer incidence rates (liver, kidney, lung, and leukemia) had been elevated in the ZIP Codes around Kelly during 1990 – 1994 as compared to the incidence rates found in the Hispanic population for the state of Texas [1].

Cancers usually involve a latency period - the period from the time of exposure or initiation until the onset and diagnosis of disease (generally 10–30 years, although some leukemia have been reported in as little as 3 years following exposure) [26]. Therefore, cancers reported during the time period examined (1990–1994) could have been the result of past exposures. Of the

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biologically plausible cancers reported, leukemia is an outcome that ATSDR is continuing to investigate (see ATSDR Health Outcome Data Evaluation Health Consultation) [25]. The incidence of liver cancer was elevated throughout a large part of south Texas, and ATSDR is continuing to investigate liver cancer [25]. The areas around Kelly AFB that had an increased incidence of kidney and lung cancer did not correspond well with the areas where the highest air contaminant levels were predicted by air dispersion modeling.

Statistically significant elevations of leukemia in three ZIP Codes (1990–1994) have been reported by the Cancer Registry Division of the Texas Department of Health. Two of the ZIP Codes are in the predominant downwind direction and the third is off-base military housing. ATSDR has investigated the elevations and distribution of leukemia types in specific ZIP Codes (see ATSDR Health Outcome Data Health Consultation)[25]. Of the chemicals known to have been emitted by Kelly AFB, benzene and 1,3-butadiene are of concern because both have been associated with leukemia in epidemiological studies of workers and because the bone marrow is a target organ for both chemicals in animal studies [27–30](see Appendix C).

The limited available data are inadequate for conducting a comprehensive assessment of potential cumulative exposures to past air emissions. Assessments of available data do not indicate a public health concern but these data are incomplete and contain more uncertainty than data collected under regulatory agency oversight (e.g. State and Federal programs under the Clean Air Act). Because of the magnitude of uncertainty and because biologically plausible health outcomes were reported in areas where people may have been exposed, ATSDR concludes that further analysis of cancer health outcomes should be performed. This further analysis is found in the Health Outcome Data Evaluation Health Consultation [25].

Susceptible Populations

ATSDR reports information on populations that may be of special interest and site-specific activities addressing potentially susceptible populations.

Children (Child Health Considerations)

Children may be at increased or decreased risk from chemical exposures. Factors that may affect their susceptibility include activity patterns, pharmacokinetic processes (ventilation rates, metabolism rates, and capacities), or pharmacodynamic processes (toxicant-target interactions in the immature hematopoietic system) [31].

Infants and children may be more vulnerable to leukemogenesis because the hematopoietic cell populations are differentiating and undergoing maturation. No data from human studies were found to indicate that children are more sensitive to benzene toxicity than are adults. Some studies have associated acute nonlymphocytic (myelocytic) leukemia and parental occupational exposures to benzene [32]. In children, the predominant type of leukemia is lymphocytic, while in adults, a combination of myeloid and lymphoid is predominately found [33]. Recent evidence suggests that *in utero* exposures may lead to leukemia [34].

Gender

No human exposure data were found to indicate that benzene affects human males and females differently.

Genetics

Individual risk factors influence an individual's unique tolerance or susceptibility to exposure and progression to disease. Polymorphisms are variations between individuals' genetic makeup which can result in changes in the way an individual responds to chemical exposures. While the genetic makeup of each individual is unknown, research indicates that certain variations in genetic makeup can account for differences in the way an individual responds to exposure to specific chemicals. Following are examples of research that illustrate the degree of variation that may exist in a population.

Individuals lacking an enzyme involved in the detoxification of a benzene metabolite could be susceptible to benzene toxicity. The lack of this enzyme appears to result from a true polymorphism in the NQO1 gene with a frequency of 13% in a reference population [35].

CYP2E1 activity in human hepatic microsomes has been shown to vary by 13-fold [36]. Differences in CYP2E1 between individual humans could indicate potential differential susceptibility to benzene and 1,3-butadiene toxicity.

Asthmatics

Individuals sensitive to respiratory irritants may experience respiratory effects at levels below where non-sensitive individuals experience respiratory effects.

Summary

While risk factors such as rates of genetic polymorphisms and asthma are not known for this population, developing hematopoietic systems may be more susceptible to insult from volatile organic compounds such as benzene and butadiene.

Conclusions and Recommendations

1. Individual contaminants from stationary sources.

Air dispersion modeling indicates that stationary source emissions were <u>unlikely</u> to have resulted in off-base exposures to individual chemicals at levels of public health concern and present *no apparent health hazard*.

Hexavalent chromium air emission data (before 1980) submitted by Kelly AFB are not sufficient for ATSDR to make a determination of public health significance and therefore represent an *indeterminate health hazard*.

Recommendation: Further investigate potential past air emissions of hexavalent chromium from Kelly AFB or include plausible health outcomes in the proposed mortality study (Kelly AFB Civilian Worker Mortality Study) to be conducted by the Kelly AFB Health Issues Working Group.

2. Individual contaminants from aircraft sources.

Air dispersion modeling indicates that aircraft emissions of JP-4 jet fuel were <u>unlikely</u> to have resulted in off-base exposures to individual chemicals at levels of public health concern and present *no apparent health hazard*.

3. Chemical mixtures from stationary and aircraft sources.

The uncertainty in potential interactions from exposure to the chemical mixture represents an *indeterminate health hazard*. Statistically significant elevations in leukemia have been previously reported in downwind ZIP Codes and off-base military housing.

Recommendation: Further investigation of elevated leukemia outcomes. This recommendation has been addressed by the Division of Health Studies, ATSDR, and reported in the Health Outcome Data Evaluation Health Consultation.

4. Air dispersion modeling sensitivity analysis suggests that selection of input parameters such as building height, building downwash, landscape type, and combinations of these parameters could result in higher estimates of on-base contaminant concentrations, but will not affect off-base concentrations. These conclusions are based in part on emissions inventory data and estimated air concentrations from air dispersion modeling. The uncertainties of these data are discussed in this report and considered in these conclusions.

Recommendation: Consider biologically plausible health outcomes from potential on-base exposures in the proposed mortality study (Kelly AFB Civilian Worker Mortality Study) to be conducted by the Kelly AFB Health Issues Working Group.

5. Data are not available for the evaluation of misting or the incineration of cyanide waste.

Authors:

David A. Fowler, PhD Consultations Section Exposure Investigations and Consultations Branch Division of Health Assessments and Consultations

Brian Kaplan, MS

Consultations Section

Exposure Investigations and Consultations Branch Division of Health Assessments and Consultations

Reviewers:

John E. Abraham, PhD, MPH Chief, Exposure Investigations and Consultations Branch Division of Health Assessments and Consultations

Susan Moore

Chief, Consultations Section Exposure Investigations and Consultations Branch Division of Health Assessments and Consultations

Ken Orloff, PhD Acting Associate Director of Science Division of Health Assessment and Consultation Appendix A Air Dispersion Modeling Methodology

Air Dispersion Modeling

This appendix presents ATSDR's rationale for the use of models to estimate the concentration of ambient air pollutants from past operations at Kelly Air Force Base.

Air Modeling

Air dispersion models are mathematical equations that *predict* (simulate or model) the movement of chemicals in the air. This movement is also called dispersion since the chemicals disperse after they are released into the air. The mathematical equations are entered into a computer program for ease of use. Data needed for these air dispersion models include weather data, the amount of pollutants released to the air over time, site topography, and site geometry. In studies comparing estimated concentrations from air dispersion modeling to air sampling measurements, the Industrial Source Complex (ISC) model used here, in areas similar to Kelly Air Force Base have been shown to be accurate within two-times to one-half the actual result [7]. For instance, if the "real" value is 1 μ g/m³, the model could show a range of 2 μ g/m³ to 0.5 μ g/m³. The largest uncertainty is the emission data which are not accounted for in this error range.

Where air monitoring shows a "real" result for a snapshot in time on one specific location, the model produces one result for each hour modeled at each specified location that must be adjusted for this error range. The modeled hourly results can be used to calculate 24-hour or annual averages or maximums.

Limitations of air models also include:

- Difficulties in obtaining representative meteorological data and emissions data.
- Large uncertainties at short time frames such as one hour or one day. Models are better at predicting long term averages such as one year.
- Complex meteorological and terrain conditions that are not accounted for in the meteorological data and the mathematical equations.
- Results that are approximations with some models validated in the field.

Four advantages of models:

- Models can be used to estimate a substance's concentration for different time periods for which both emissions and meteorological data exist. The ISCST3 model used in this report generates an hourly model. The hourly results can be compiled to generate maximum and average values. Maximum and average results can also be generated for any time period such as a day, month, or year.
- Models can be used to estimate the level of various substances existing in the ambient air as a result of emissions from a single source or multiple sources.
- Models can average short-term fluctuations in emissions and meteorological conditions, resulting in a long-term average.
- Models can estimate a substance's concentration at an unlimited number of locations.

Air Sampling

Air sampling using conventional equipment has the advantages of producing data that are

considered "real" results. "Real" in the sense that the mix of chemicals identified actually existed in the air at the location and time the sample was taken. Moreover, this mix of chemicals was the result of many different sources. Conventional equipment is defined here as fixed stationary samplers with samples collected by drawing air through a filter or tube and the filter and tube analyzed at a later time for the chemicals collected. Although the sample is considered "real," there are several disadvantages in the sampling procedure:

- Sampling substances arising from many and varied sources hinders the correlation of an air sample to a single facility. Sources not pertinent to the investigation could influence the interpretation of the results. For instance, air samples collected near idling buses may have higher concentrations of chemicals found in diesel exhaust than is present in ambient air as a result of emissions from the source being investigated.
 - Sampling results are based on conditions at the time of the sampling event. These conditions include the meteorological conditions and the amount and rates at which the chemicals were released. These conditions could be an extremely low or high condition and not representative of average conditions. Conversely, samples are usually collected over a period of time (several hours to 12 hours), consequently, the result would average out short term small and large transient chemical concentrations.
- Air sampling is expensive and takes a long time to obtain representative results.

Air Modeling Input Parameter Comparison

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Peer reviewers of the Phase I PHA recommended ATSDR investigate the effect of different input values for the half life of hexavalent chromium [37]. This section describes the results of variables in that and other input parameters. Table A-1 provides a summary of these details.

ATSDR investigated various input parameters to determine their potential effects on the results of air dispersion modeling. The following issues related to air dispersion modeling were addressed using the U.S. EPA Industrial Source Short Term (version 3, ISCST3) gaussian dispersion model:

- 1. effect of using different half lives for hexavalent chromium. 1 to 2,160 minutes was used
- 2. building downwash (effect with and without)
- 3. landscape (effect of using rural or urban)
- 4. building height (20 or 32 feet)
- 5. combinations of parameters

Five years of meteorological data were used (January 1986 through December 1990) from the San Antonio International Airport for surface air and the Del Rio International Airport, Del Rio Texas for upper air data.

For the analysis of the different parameters, ATSDR assumed 12 chromium emission points. The details of these emission points are presented in Appendix B, Attachment 1.

Results are depicted in Figure A-1 and Figure A-2. While the true concentration is unknown (actual emission rates are not known), Figure A-1 indicates that as the half-life of hexavalent chromium increases, the estimated concentration at a downwind point near the base perimeter increases until the half-life value used is about 3 - 4 hours, at which point the estimated concentration is not changed by the longer half life. Use of a half-life above 3 - 4 hours does not increase the concentration at the model locations at the base perimeter. However, the distance from the base where the maximum concentration could be found would increase as the half-life is increased above 3 - 4 hours.

Figure A-1. Hexavalent Chromium Half-lives. Results indicate that chromium concentration at a point in the community near the fenceline north of the base^{*} increases as the half-life approaches 3 - 4 hours (180–240 minutes). The concentration then becomes stable.



Figure A-2 indicates that the use of rural or urban landscape, with or without downwash, at a 20 or 32 foot release height will not result in a significant change in the estimated concentration of hexavalent chromium off base. On base concentrations are most significantly affected by release height, rural landscape, and without downwash, respectively, resulting in the highest breathing zone concentrations nearer the source. Because the meteorology used for other contaminants is the same as that used for hexavalent chromium, the relative concentration differences would also apply to all other contaminants. This suggests that, depending on the input parameters selected, the concentration of contaminants on base could vary by a factor of about 3 at a specific point within 300 meters of the source and a factor of about 50 depending on the receptors location inside the base boundaries. Figure A-2 also illustrates that selection of input parameters will <u>not</u> affect the off-base concentrations of contaminants from Kelly AFB.

^{*}This point is located at 641,600 meters west and 4,173,700 meters north of the origin of the geographic statewide grid, Texas South Central Zone, North American Datum of 1983.



Figure A-2. Input Parameter Comparison. Selection of model parameters shown in the Figure have no effect on off-base concentrations of contaminants, but may have significant effects upon on-base concentrations.

Table A-1. Summary of parameters and assumptions common to aircraft and stationary source modeling.

Category	Data and Assumptions					
Dispersion Modeling	Used the U.S. EPA Industrial Source Complex Short Term Version 3 (ISCST3) model. Assumed that deposition or degradation in the atmosphere did not occur. The rural dispersion parameters were used. Downwash was not used except in the analysis of the parameters in Appendix A.					
	The model assumptions inherent in the ISC model were used and include the Gaussian dispersion parameters derived from Pasquill-Gifford. These assumptions and other background data on this model can be found in two U.S. EPA manuals (Vol. I and Vol. II., 1995, <u>http://www.epa.gov/scram001/tt22.htm#isc</u> .					
	See Appendix B for additional details.					
Meteorological data	Five years of meteorological data were used (January from the San Antonio International Airport for surfac Airport, Del Rio Texas for upper air data.	1986 thro e air and t	ugh December 1990) he Del Rio International			
Aircraft Modeling						
Dispersion Modeling	The ISC model input parameters were set so the emis volume sources. The size of the volume and its locati specific values. The sensitivity of these values to the volume sources were assumed to disperse due to meter operations may have created additional dispersion wh downwind concentrations. See Appendix B for additional	sions were on behind final result corologica ich could onal detail	e contained in a series of the plane were set to ts was not tested. The l conditions. Aircraft lead to lower s.			
Number of Operations	336,000 total with 168,000 takeoffs and landings	The nun peak val on the re AFB em previous or highe assumed one land consistir shutdow approacl	ber of operations is a ue in 1964 and is based elocation of a Kelly ployee.* Operations in years could be lower r. An operation to be one takeoff and ing and each operation of taxi, startup or n, runway rollway, and o or takeoff.			
Emissions	Based on engine tests of the B52H engine (TF33-3) and the F16	Other B: Pratt and Previous models of may hav emission compare engine e different condition emission	52 engines included the 1 Whitney J-57. 5 engines or different 5 f the same engine type e had more or less is. The tested engine d to actually used missions may be due to wear or load is. Assumed that all is were gaseous. See x B for details			
Time in mode	Based on a USAF reference. See Appendix B for deta	ils.				
Location	Based on existing runway and main long taxiway.		A second runway that no longer exists was			

 Table A-1. Summary of parameters and assumptions common to aircraft and stationary source modeling.

Category	Data and Assumptions				
Stationary Source Modeling					
Location and rate of releases	Data is Appendix B, Attachment 1 and based on three Air Force reports (March 27, 2000, June 2000, and December 2000). Emissions reported on an annual basis was averaged over a year.				
Stack parameters	Based on March 27, 2000, June 2000, and December 2000 Air Force reports. Sources with no data were assumed to have a building height of 6.2 meters, exit gas temperature of 20°C, exit gas velocity of 0.1 meters/second, and an inside diameter at the release point of 1 meter. These values are conservative in that they produce higher ambient air concentrations.				

* Aircraft Noise in the Vicinity of Kelly Air Force Base, San Antonio, Texas (Updated Report), March 1994. HMMH Report No. 292610-B. Prepard fro U.S. Department of Justice by Harris, Miller, Miller & Hanson Inc., Lexington, Massachusetts. Appendix B Stationary and Aircraft Emissions

Appendix B

This appendix presents ATSDR's approach to estimating the concentration of ambient air pollutants from past operations at Kelly Air Force Base (AFB). The approach was completed in two steps: estimating emission rates and modeling the dispersion of the emissions. These steps are addressed individually for stationary source (industrial) emissions and aircraft emissions.

Air dispersion models are mathematical equations that *predict* (simulate or model) the movement of chemicals in the air. This movement is also called dispersion since the chemicals disperse (spread out) after they are released into the air. The mathematical equations are put into a computer program for ease of use. Data needed to estimate emissions rates include weather data, adjacent land use, building height and size, the amount of pollutants released to the air over time, and the release location of the pollutants. More specifically, the data needed include:

- Temperature of exit gas
- Diameter of stack at exit
- Exit gas velocity
- Location of the release in geographic coordinates
- Amount of pollutant being released over time (rate of release)
- Release height or stack height

Stationary Source Emissions

ATSDR obtained the location and the rate of releases from Kelly AFB (see Appendix B, Attachment 1). The information was provided in a report dated March 27, 2000, and updated in June and December, 2000. Where known, this information included building numbers and heights, heights of vents or stacks, descriptions of processes, specific chemicals, usage and emissions estimates, assumptions, and sources of the information.

ATSDR compared the past emission rates with previously modeled current emission rates to determine the need for additional modeling of those chemicals. ATSDR also considered whether stack heights, building locations, or other parameters were different and therefore, suggested whether additional modeling would be necessary. The following chemicals were addressed as stationary source emissions of potential concern with results presented in Table B-1: tetrachloroethylene (PCE), hexavalent chromium, methylene chloride, methyl ethyl ketone, benzene, ethyl benzene, formaldehyde, toluene, and xylene. Building numbers, locations, and emission values are listed in Attachment 1 for these chemicals. ATSDR was not able to obtain the temperature of the exit gas, diameter of stack at exit, exit gas velocity, or the release height for each of these release points. As a result, ATSDR assumed a set of stack values that would overestimate ambient air concentrations. These values are:

- Building Height: 6.200 meters (approximately 20 feet)
- Stack Gas Exit Temperature: 293 degrees Kelvin (20°C or 68°F)
- Stack Gas Exit Velocity: 0.10 meters/second (0.33 feet/second)
- Stack Inside Diameter at Release point: 1 meter (3.2 feet)

ATSDR assumed that all modeled chemicals were in the gas form and deposition (dry or wet) was not occurring. This assumption can overestimate the amount of chemical in the air. The likely form of most metals and hexavalent chromium in the air is as an aerosol or absorbed onto particulates. The amount of deposition of the aerosols and particulates would be a function of their size and mass distribution. These properties were not known so ATSDR assumed all the chemicals including chromium was in the gas form.

ATSDR also assumed that the nearby land use was rural and that the building height and size where the release occurred and nearby buildings did not influence the dispersion of the chemicals (this influence is called building downwash). In Appendix A, Air Modeling Input Parameter Comparison, ATSDR verified these inputs to be producing higher predicted concentrations than would occur if other inputs were used.

ATSDR modeled these emissions estimates using EPA's ISCST3 model. The ISCST3 model in flat terrain, as in the case of Kelly AFB, as an uncertainty from ½ to 2 times the predicted concentrations [7].

The predicted concentrations indicate that non-cancer health effects would be unlikely as no contaminant concentration exceeded noncancer comparison values (Table B-1). Two chemicals exceeded cancer comparison values and were further evaluated: tetrachloroethylene (PCE) and methylene chloride (Table B-1). Hexavalent chromium emissions were not included because the data provided were insufficient (see Hexavalent Chromium below for an explanation). Following are discussions of the emissions of these chemicals and health implications of the estimated levels.

Hexavalent chromium

The emissions from the plating operations were expected to be the largest potential contributor of hexavalent chromium emissions. ATSDR evaluated two time frames relating to hexavalent chromium emissions. Hexavalent chromium was emitted from five plating shops. The most significant were located in Buildings 258/295 and Building 301. Buildings 258/259 began operation in 1942 and shutdown in 1977. Building 301 replaced Building 258/259 in 1977 and included scrubbers to control emissions. The emission rates of hexavalent chromium from Building 258/259 are not known. The emission rates from Building 301 are based on stack tests completed in 1980. The time prior to the 1980 stack test is used to define past exposures because of the unknown emission rates from Building 301 prior to 1980 and unknown emission rates from Buildings 258/259.

ATSDR was interested in the time frames before and after the Building 301 stack test. Emissions before 1977 would potentially represent higher concentrations because it is not known if pollution control was in use in Building 258/259 while emissions from 1977 to 1980 are not certain. Emissions from 1980 would be more certain based on the stack test and be similar to current emissions. Emissions from 1980 would potentially represent lower concentrations than prior to 1977. Data located for chromium plating emissions were rare (3 data sources). Hexavalent chromium emissions were evaluated for 3 scenarios based on data sources and timeframes related to changes in emissions:

(1) air emissions estimated from *chromic acid usage* data before 1980,

(2) air emissions estimated from *energy usage* before 1980, and

(3) air emissions measured from stack emissions data and applied to 1980 and after.

Chemical	Estimated emission (TPY) *	Estimated Concentration ^b (µg/m³) ^c	Chronic Non-Cancer Comparison Value (µg/m ³)	Worker Exposure Levels (µg/m ³)*	Cancer Comparison Value (µg/m³)	Estimated Cancer Risk
hexavalent chromium	NA	-	_	-	-	-
methylene chloride	2940	123	1060 °	87,000	3 ^f	5E-05
PCE	1490	142	271 °	678,000	3 ^g	7E-05
benzene	0.04	0.00046	13 ° intermediate	320	0.1 ^f	5E-09
formaldehyde	0.13	0.00075	10 ^e	922	0.08 ^f	1E-08
methyl ethyl ketone	305.	13	1000 °	1500	-	NA
toluene	0.0116	0.00013	300 °	375,000	-	NA
xylene	0.00517	0.00006	434 °	435,000	-	NA
ethyl benzene	0.0175	0.0002	1000 h	435,000	-	NA

Table B-1. Estimated maximum past ambient air concentrations from stationary source emissions.

a TPY or tons per year

b The estimated concentration was determined as the maximum off-base concentration.

c $\mu g/m^3$ or micrograms per cubic meter

d intermediate exposure value of <1 year used as there is no chronic value.

e ATSDR minimum risk level (MRL)

f ATSDR cancer risk evaluation guideline (CREG)

g EPA Risk Based Concentration, Region 6.

h EPA RfC or EPA Reference Concentration

NA Not Applicable or Not Available

NIOSH values were used for all except PCE and formaldehyde, where OSHA values were used. Values represent up to 10-hour daily exposures for up to 40 hours/week. OSHA values represent an 8 hour day, 5 days/week.

Kelly AFB provided two data sources for estimating hexavalent chromium air emissions prior to 1980 (scrubbers were installed in 1980 for pollution control). The data sources were

(1) *usage of chromium trioxide* (used to make chromic acid) from one year of operation during the 1980s (the specific year was not reported), and

(2) *emission estimates from energy usage* during one test of stack emissions on one scrubber (pre-scrubber concentrations were estimated from Tinker AFB energy use data formulation).

Kelly AFB provided a third source of information from *measured stack emissions* from one test of one scrubber in 1980 in Building 301. The quality of these data obtained from analysis of scrubber stack emissions is much higher than for either the chromic acid usage data or the energy usage data. Data from chromic acid usage and energy usage contain data gaps and little corroborating information, which ATSDR deems essential to reduce uncertainty in the results. Although the quantity of stack emissions data is low, ATSDR acknowledges these data are probably indicative of emissions after scrubbers were tested in 1980 in Building 301. Corroborating evidence from current emissions data supports this judgment. Sufficient quantitative data have not been provided for a quantitative assessment of emissions after the scrubber were installed in 1980. While a quantitative assessment is not possible, a qualitative evaluation does not indicate that evidence exists indicating levels of public health concern were likely to have been present after the scrubbers were tested (1980). ATSDR acknowledges that any evaluation would contain much uncertainty and insufficient to make a public health call.

Some of the information ATSDR requested from Kelly AFB could not be located. This information includes the number of air emissions scrubbers, when the operation began, size of chromic acid baths, chromic acid strength, and electricity used in plating operations at Kelly AFB. Because of these unknowns, it is not possible for ATSDR to estimate concentrations from potential past air emissions of hexavalent chromium with an appropriate degree of confidence to draw conclusions related to past exposures. Therefore, ATSDR concludes that the health hazard to the community before scrubbers were installed is *indeterminate*.

Methylene chloride

Non-cancer health effects from exposures to methylene chloride would not be expected because the maximum off-base concentration of methylene chloride did <u>not</u> exceed non-cancer comparison values (see Table B-1). The maximum off-base concentration of methylene chloride exceeded a cancer comparison value. Therefore, methylene chloride was further evaluated to determine the estimated risk for developing cancer from the maximum exposure. Methylene chloride is considered a probable human carcinogen based on inadequate human data and sufficient animal data. The estimated risk for a continuous lifetime exposure at the maximum concentration is considered a *low increase* in risk. Animal studies served as the basis for calculating risk as no human cancers have been reported in the scientific literature <u>at these</u> <u>estimated levels</u>. Using the maximum value estimated air concentrations based on the modeling, risks are likely to be overestimated. Using the maximum estimated concentration in the community, potential exposures levels are about 3500 times less than levels potentially associated with reported cancer effects in humans [38]. Although there is some risk from exposure to methylene chloride, ATSDR would not expect that an increase in cancer would be observed in the community from exposures to these estimated levels of methylene chloride.

PCE

The maximum off-base concentration of PCE did not exceed non-cancer comparison values (see Table B-1). While PCE has been confirmed as an animal carcinogen, the carcinogenicity of PCE in humans continues to be investigated. The maximum off-base concentration of PCE exceeded a cancer comparison value based on animal studies and was further evaluated. The estimated risk for developing cancer in the community from a continuous lifetime maximum exposure to the maximum concentration of PCE is considered a *low increase* in risk. PCE exposure <u>at these levels</u> has not been associated with cancer in humans. ATSDR would not expect adverse health effects would be observed as a result of exposures at these levels under these exposure conditions [39].

Aircraft Emissions

ATSDR estimated the concentrations of organic chemicals in the ambient air from aircraft emissions using air modeling. Data on metal emissions were not available. This section discusses the inputs used in the model and the modeling process. ATSDR reviewed data from the Air Force and Navy on airplane emissions to select model input parameters. The input parameters were selected to be conservative (i.e., worst emissions) in most cases. As a result, ATSDR modeled the maximum reported annual operations of 336,000 in 1964 and assumed all operations were conducted by the B52H aircraft which emits the most pollution overall from data ATSDR reviewed. ATSDR also modeled emissions from a F16 aircraft to provide perspective. A B52 has eight engines and an F16 has one engine. Emissions information is available on 69 organic chemicals and ATSDR modeled the emissions of six chemicals based on amounts emitted and toxicity. These chemicals included acetaldehyde, acrolein, benzene, butadiene, formaldehyde, and naphthalene. The concentrations of these six chemicals in the environment were estimated at 5,100 point locations (the points were 300 meters apart) in and around Kelly AFB. Specific details of the modeling are presented in the remainder of this section.

Model Inputs

The Industrial Source Complex-Short Term (ISCST) model was used to perform the air modeling. To use this model, information on the source of pollutants, ambient meteorology, and information on receptor locations must be entered into the model. The model simulates the movement of the pollutants in the atmosphere and calculates a concentration at the given receptor locations. *The emissions were treated as a series of volume sources behind the aircraft (see page 32 for details)*.

Source of Pollutants

The source of the aircraft emissions was aircraft operations at Kelly AFB. To use the model, ATSDR must know the amount of each type of pollutant released per unit of time and the location of the release. Since the aircraft move throughout the base, the release of the pollutants would occur at many different locations.

Obtaining information about the source and location of pollutants from the aircraft was a four step process:

- 1. Determining the types of aircraft at Kelly AFB
- 2. Obtaining the number of flight operations performed by each aircraft (takeoffs, landings, others)
- 3. Obtaining the amount and types of pollutants released from each aircraft.
- 4. Identifying the movement of the aircraft (location and time spent at the location).

Types of aircraft at Kelly AFB.

Kelly AFB was established as a military air field in 1917 and many different types of aircraft have flown through Kelly AFB. ATSDR obtained the following list of aircraft from Kelly AFB [40, 41]:

					,	
A-10	B-58	C-141	C-97	F16	KC-135	T-34
A-4	B-727	C-17	DC-9	F-18	P-38	T-37
A-6	B-757	C-2	F-100	F-4	P-47	T-38
B-1	C-118	C-21	F-101	F-80	P-51	T-39
B-17	C-119	C-47	F-102	F-84	SR-71	T-41
B-24	C-12	C-5	F-104	F-86	SW-4	T-43
B-29	C-121	C-54	F-105	F-89	SWB	T-45
B-36	C-123	C-7	F-106	F-94	T-1	T-6
B-50	C-124	C-74	F-14	FB-111	T-28	UH-1
B52	C-130	C-9	F-15	KC-10	T-33	XC-99

ATSDR limited the aircraft emissions modeling to the B52H because it was the largest emitter of pollutants in the limited available data. This simplification is a conservative assumption in that ATSDR modeled a worst-case (most emissions) scenario. This simplification is explained in additional detail in the following sections. Because the operations per type of plane used in 1964 was not known to ATSDR at the time this report was prepared, ATSDR also modeled emissions from the F16 to simulate a scenario with lower emissions. The F16 did not exist in 1964 but is used as a surrogate for a low emission military aircraft.

Number of flight operations performed by each aircraft

Ideally, the number and types of operations for each aircraft is used. ATSDR only found historical information on the total number of operations per year as depicted in Figure B-1. An aircraft operation is one take off or one landing. A touch and go (landing and immediate take off) is two operations [14].





To present a worst-case (most emissions) scenario, ATSDR modeled the 1964 operations of 336,000 operations per year. Since information on all of the types of aircraft was not available, ATSDR assumed all 336,000 operations were performed by the B52H or the F16.

Amount and types of pollutants released from each aircraft ATSDR found emissions data on the following aircraft:

A-10A	F-106 (A,B)	T33A
B-52 (D, F) B-52H	F15 (A,B,C,D)	T-37B
C-130 (A, D, E, H)	F-16 (A,B)	T-38 (A,B)
C-141 (A,B)	F4 (C, D, E, F)	T-39 (A,B)
C-21A	FB-111A	T-41 (A,B,C)
C-5 (A,B)	KC-10A	T-43A
C-9A	KC-135 (A,D)	

Available emission data included carbon monoxide, carbon dioxide, total hydrocarbons, nitrogen oxides, particulate matter, and specific organic chemicals. Carbon monoxide, carbon dioxide, nitrogen oxides, and particulate matter were not modeled because these compounds are typical in urban air while total hydrocarbons is not sufficiently specific for a toxicological evaluation. ATSDR focused its modeling on specific organic chemicals which are listed below.

ATSDR obtained and used emissions from these aircraft using JP-4 fuel versus the currently used JP-8 fuel. Kelly AFB completed the conversion from JP-4 jet fuel to JP-8 jet fuel in 1994[13]. JP-4 has been used by the Department of Defense since 1951 [42].

As stated previously, ATSDR simplified the modeling to the emissions of the B52H because it presented a worst-case (most emissions) scenario. ATSDR identified the B52H as the worst case as follows. The F16 was used to simulate planes with lower emissions.

ATSDR reviewed the available emissions data by total hydrocarbons for different aircraft using JP-4. ATSDR identified the plane and engines that emitted the most hydrocarbons. Table B-2 lists these aircraft and engines by flight operation mode because the aircraft and their engines emit different amounts of chemicals during each mode.

Table B-2. Summary of aircraft and engines with the most hydrocarbon emissions per aircraft mode [43].

			Hydrocarbon Emissions Per Plane Per Event
Mode	Plane	Engines	(Metric Tons)*
Startup	B52H	TF33-3	0.0582
Taxi Out	B52H	TF33-3	0.113
Engine Check	B52H	TF33-3	0.0582
Runway Roll	B52H	TF33-3	0.000176
Climb 1	B52H	TF33-3	0.000193
Climb 2	B52H	TF33-3	0.000213
Approach 1	KC-135	J57-59W	0.00251

			Hydrocarbon Emissions Per Plane Per Event
<u>Mode</u>	Plane	Engines	(Metric Tons)*
Approach 2	C5	TF39-1	0.000783
Landing	B52H	TF33-3	0.00699
Taxi In	B52D	J-57-19W/J-57-43WB	0.0418
Shutdown	B52D	J-57-19W/J-57-43WB	0.0106
Touch&Go	KC-135	J57-59W	0.0033

* The hydrocarbon emissions per plane per event from this reference are most likely not correct. ATSDR checked these values against possible derivation. These hydrocarbon emissions per plane per event are based on time-in-mode, engine setting, and HC emission rate. The power setting for taxi-out is idle which has the highest HC emissions rate. For the TF33-3 engine at idle, the rate is 84 g/kg fuel. The fuel use rate is 0.11 kg/s so the HC emission rate is 9.24 g/s. For 113 kilograms HC emitted during taxi out (Table B-2), the time-in-mode would need to be 3.4 hours which seems very unrealistic. ATSDR checked the KC-135A from this reference for taxi-out and came up with 11.5 hours which is even more unrealistic. ATSDR suspects a systematic error in Table A of Seitchek [43]. It's possible that the units for Table A are kilograms and not metric tons. Because the values in Table B-2 were only used for a comparison among planes and not used in the emissions modeling, the error in Seitchek (1985) does not change our results. The hydrocarbon rates used in the modeling was 94 g/kg fuel 0.14 kg/s of fuel (Spicer et al 1988) [16]. These values are similar to Seitchek (1985). The times-in-mode used in the ATSDR modeling was 9 minutes for taxi-out (Naugle et al 1975) for a total of 7.1 kg HC released during taxi-out [44].

These data indicate that the B52H aircraft emitted the most hydrocarbons overall [43]. The C-5 had the highest emission rates during Approach 2 while the B52D had the highest emission rates during Taxi In and Shutdown and the KC-135 had the highest emission rates during Approach 1 and Touch and Go.

The hydrocarbon emissions for each mode of the B52H with the TF33-P3 engine (Table B-3) was compared to Table B-2. From these tables, the B52H is not the worst emitter in four of the 12 modes, with the most significant difference is in Approach 2. Because the B52H was the worst emitter for 8 of the 12 modes, ATSDR decided to use the B52H as the worst-case aircraft to model.

	Hydrocarbon Emissions Per Plane Per Event
Mode	(Metric Tons)*
Startup	0.0582
Taxi Out	0.113
Engine Check	0.0582
Runway Roll	0.000176
Climb 1	0.000193
Climb 2	0.000213
Approach 1	0.00178
Approach 2	0.000595
Landing	0.00699
Taxi In	0.0413
Shutdown	0.0105
Touch&Go	0.0029

Table B-3. Hydrocarbon emissions per event for the B52H aircraft with TF33-P3 engines.

*USAF Aircraft Engine Emissions Estimator, Glenn D. Seitchek, ESL-TR-85-14, November 1985.

Additional parameters required for modeling the B52H/TF33-P3 are time in mode, fuel flow, and hydrocarbon emissions per fuel rate shown in Table B-4.

				HC Emissions
Aircraft Mode	Engine Thrust	Minutes*	Fuel Flow Per Engine (1000 lb / hr)**	Per Engine (lb/1000 lb fuel)**
Startun	Idle	20	(1000 ID / III) 1 052	94.00
Startup	Iule	20	1.052	94.00
Outbound Taxi	Idle	9	1.052	94.00
Engine Check	Military	4.5	7.105	0.03
Runway roll	Military	0.7	7.105	0.03
Climbout I	Military	0.7	7.105	0.03
Climbout II	Military	0.8	7.105	0.03
Approach I	Idle	3	1.052	94.00
Approach II	Idle	1	1.052	94.00
Landing on	Idle	1	1.052	94.00
runway				
Inbound Taxi	Idle	12	1.052	94.00
Idle at shutdown	Idle	4.8	1.052	94.00

Table B-4. Operating parameters for a TF33-P3 engine in different B52H aircraft modes.

*USAF Aircraft Pollution Emission Factors and Landing and Takeoff (LTO), Dennis Naugle, et al, AD/A-006 239 (February 1975)

**Aircraft Emissions Characterization," C.W. Spicer, M.W. Holdren, S.E. Miller, D.L. Smith, R.N. Smith, D.P. Hughes. Final Report, March 1988, Engineering and Services Laboratory, Air Force Engineering & Services Center, Tyndall Air Force Base, ESL-TR-87-63.

The constituents of the total hydrocarbons (HC) from a TF33-P3 engine have been reported by

the Air Force and shown in Table B-5. The Air Force reports the emission test results in $\mu g/m^3$ for polyaromatic hydrocarbons and parts per million Carbon (ppmC) for all other pollutants. $\mu g/m^3$ and ppmC are converted to percent weight of hazardous air pollutant (HAP) emitted in the hydrocarbon emissions using the following formula:

% wt HAP/HC = ([HAP] / [HC]) x (Number of C in HC / Number of C in HAP) x (MW_{HAP} / MW_{HC}) where:

 $[HAP] = \text{concentration of organic compound in ppm}_{C}C$ $[HC] = \text{concentration of total hydrocarbons in ppm}_{C}C$ Number of C = Number of carbon molecules = 9.3* $MW_{HAP} = \text{Molecular Weight of the HAP}]$ $MW_{HC} = \text{Molecular weight of the total hydrocarbons} = 130*$

*Douglas, Everett, Naval Aviation Depot, Naval Air Station, San Diego, California. Email record of personal communication regarding information about converting units and data on the number of carbons and molecular weights of total hydrocarbons in jet fuel, February 12, 2001. A derivation of this formula is presented in Response to Comments.

From this data, the amount of HAPs emitted per unit time (e.g., grams/second) is calculated, in general, as follows:

Amount of fuel burned per time in each mode * Amount of HC emitted per fuel burned * Number of engines * % wt HAP/HC * Time in mode * Number of operations per hour = Amount of HAPs emitted per unit time (grams/second) * multiplication

Table B-5. Chemicals in exhaust from the TF33-P3 engine using JP-4 jet fuel.

	Power Setting								
	ldle		30	30%		75%		100%(military)	
		%wt		%wt		%wt		%wt	
	ppm C	HAP/HC	ppm C	HAP/HC	ppm C	HAP/HC	ppm C	HAP/HC	
Methane	37.57	0.03425	2.425	0.03578	0.704	0.13454	0.439	0.15851	
Ethane	5.098	0.00435	0.334	0.00462	0.01	0.00179	0.004	0.00135	
Ethylene	47.405	0.03777	15.171	0.19568	1.368	0.22854	0.164	0.05176	
Propane	0.893	0.00075	0.026	0.00035	0.005	0.00088 ·	<0.001		
Acetylene	28.368	0.02098	3.925	0.04698	0.374	0.05798	0.068	0.01992	
Propene	43.344	0.03454	5.048	0.06511	0.289	0.04828	0.041	0.01294	
1-Butene	18.489	0.01473	1.814	0.02339	0.107	0.01787	0.049	0.01546	
1,3-Butadiene	11.981	0.00920	0.571	0.00710	0.024	0.00387 •	<0.001		
1-Pentene	5.818	0.00464	0.595	0.00767	0.065	0.01086	0.042	0.01326	
C5-ene	2.563	0.00204	0.225	0.00290	0.017	0.00284 •	<0.001		
n-Pentane	4.464	0.00366	0.112	0.00149	<0.001	•	<0.001		
C5-ene	1.584	0.00126	0.013	0.00017	<0.001	•	<0.001		
C5-ene	0.835	0.00067	0.052	0.00067	<0.001	•	<0.001		
2-Methylpentane	11.059	0.00902	0.192	0.00254	0.005	0.00086 •	<0.001		
3-Methylepentane	8.438	0.00689	0.335	0.00442	0.006	0.00103 •	<0.001		
1-Hexene	5.587	0.00445	0.521	0.00672	0.023	0.00384 •	<0.001		
n-Hexane	14.688	0.01198	0.339	0.00448	0.005	0.00086 •	<0.001		
Methylcyclopentan e + unk	7.834	0.00000	0.246	0.00000	0.059	0.00000	0.077	0.00000	
Benzene	12.499	0.00924	1.698	0.02032	0.16	0.02481	0.029	0.00849	
2-Methylhexane	25.488	0.02073	0.571	0.00752	0.012	0.00205	0.004	0.00129	
3-Methylhexane	20.534	0.01670	0.589	0.00775	0.009	0.00153	0.006	0.00193	

				Pow	er Setting			
	Idle 30% 75%			%	100%(military)			
	-	%wt		%wt		%wt		%wt
	ppm C	HAP/HC	ppm C	HAP/HC	ppm C	HAP/HC	ppm C	HAP/HC
n-Heptane	26.87	0.02185	0.606	0.00798	0.016	0.00273	0.011	0.00354
	31.824	0.02536	0.643	0.00829	0.014	0.00234	0.007	0.00221
1 Oluene 2 Methylheptene	23.27	0.01740	1.440	0.01750	0.076	0.01191	0.022	0.00652
2-Methylheptane	21.220	0.01721	0.339	0.00445	0.013	0.00221	0.009	0.00269
p-Octopo	29.015	0.02300	0.007	0.00902	0.010	0.00306	0.01	0.00321
Ethylehenzene	5 5 5 8	0.02345	0.090	0.00781	0.014	0.00238	0.009	0.00209
m+n-Xvlene	30 787	0.00413	1 332	0.00390	0.017	0.00209	0.000	0.00233
Styrene	11 174	0.02020	0.38	0.00455	0.040	0.00730	0.024	0.00351
oXviene	9,734	0.00734	0.413	0.00504	0.016	0.00253	0.008	0.00239
n-Nonane	22.406	0.01814	0.464	0.00608	0.012	0.00204	0.01	0.00321
p-Ethyltoluene	8.352	0.00634	0.346	0.00425	0.009	0.00143	0.016	0.00481
1,2,4-	15 504	0.01100	0 566	0.00005	0.040	0.00054	0.019	0.00544
Trimethylbenzene	10.001	0.01162	0.000	0.00695	0.016	0.00254	0.016	0.00541
n-Decane	21.715	0.01755	0.498	0.00652	0.019	0.00322	0.014	0.00448
Methlybenzaldehyd	8 179	0 00624	0 497	0.00613	0.059	0 00943	0.062	0.01872
e+C10H14		0.00021	0.407	0.00010	0.000	0.00010	0.002	0.01012
Undecane	26.179	0.02113	0.606	0.00792	0.027	0.00457	0.042	0.01343
Naphthalene	10.138	0.00738	0.395	0.00465	0.035	0.00534	0.05	0.01442
Dodecane	29.261	0.02360	0.522	0.00681	0.023	0.00389	0.066	0.02108
Tridecane	21.398	0.01724	0.452	0.00589	0.034	0.00574	0.08	0.02553
Formaldebyde	0.011 15.54	0.00403	0.405	0.00528	0.041	0.00092	0.104	0.05607
Acetaldebyde	10.04	0.02000	4.009	0.11000	0.423	0.15120	0.003	0.05007
Acrolein	1.002	0.00220	0.504	0.03100	0.211	0.00000	0.030	0.01704
Pronananldehyde	0.461	0.00155	0.001	0.00001	0.001	0.00133	0.001	0.00218
Acetone	<0.401	0.00001	0.200	0.00769	0.013	0.00400	0.000	0.00210
Benzaldehyde +	0.000		0.102	0.00100	0.007	0.01010	0.020	0.01000
unk	3.9303	0.00338	1.668	0.02325	0.2	0.03611	<0.001	
Glyoxal	1.68	0.00277	1.368	0.03650	0.126	0.04354	0.024	0.01567
Methyiglyoxal	5.31	0.00725	0.817	0.01804	0.077	0.02203	0.032	0.01729
Biacethyl	0.542	0.00066	0.257	0.00509	0.024	0.00615	0.013	0.00630
ĸ			Polyaror	natic Hydroca	rbons	750/		4000/
				30%		/5% 9/114		100%
	µg/m³		µg/m³		µg/m³		µg/m³	
Nanhthalene	320	0.00041	45		Q	0.00240	23	0.00116
1-methví	520	0.00041	40	0.00033	5	0.00240	2.0	0.00110
naphthalene	.430	0.00055	33	0.00068	3.6	0.00096	1	0.00050
2-methyl	250	0 000 45	40	0.004.04		0.00400		0 00055
naphthalene	350	0.00045	49	0.00101	4.5	0.00120	1.1	0.00055
Dimethyl	53	0 00007	00	0.00018	0.043	0.00001	0.064	0 00003
naphthalene	55	0.00007	0.0	0.00010	0.045	0.00001	0.004	0.00000
Dimethyl								
naphthalene	320	0.00041	33	0.00068	1.8	0.00048	0.53	0.00027
Isomer								
1,2-aimetnyi	530	0.00067	53	0.00109	3.2	0.00085	1.2	0.00060
14822 dimothul								
nanhthalene	140	0.00018	14	0.00029	0.8	0.00021	0.29	0.00015
2.6-dimethyl								
naphthalene	32	0.00004	3.3	0.00007	0.19	0.00005	0.088	0.00004
Dimethyl								
naphthalene	21	0.00003	11	0.00023	1.5	0.00040	0.1	0.00005
isomer								
Dimethyl								
naphthalene	40	0.00005	1.9	0.00004	0.72	0.00019	0.22	0.00011
Isomer	4.0	0.00004	0.00	0.00000	0.045	0.00004	0.040	0.00004
rnenanthrene	4.8	0.00001	0.22	0.00000	0.045	0.00001	0.019	0.00001

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Power Setting								
ldle		30	30%		75%		100%(military)	
	%wt		%wt		%wt		%wt	
ppm C	HAP/HC	ppm C	HAP/HC	ppm C	HAP/HC	ppm C	HAP/HC	
9.7	0.00001	0.76	0.00002	0.32	0.00009	0.11	0.00006	
8.9	0.00001	0.64	0.00001	0.27	0.00007	0.095	0.00005	
0.2	0.00000	0.012	0.00000	0.012	0.00000	0.01	0.00001	
0.2	0.00000	0.034	0.00000	0.026	0.00001	0.021	0.00001	
<0.01		<0.01		<0.01		<0.01	·	
<0.01		<0.01		<0.01		<0.01		
<0.01	·	<0.01		<0.01		<0.01		
<0.01		<0.01		<0.01		<0.01	_	
	ppm C 9.7 8.9 0.2 0.2 0.2 0.2 <0.01 <0.01 <0.01 <0.01	Idle ppm C HAP/HC 9.7 0.00001 8.9 0.00001 0.2 0.00000 0.2 0.00000 <0.01 <0.01 <0.01 <0.01	idle 33 %wt ppm C HAP/HC ppm C 9.7 0.00001 0.76 8.9 0.00001 0.64 0.2 0.00000 0.012 0.2 0.00000 0.034 <0.01 <0.01 <0.01 <0.01 <0.01 <0.01 <0.01 <0.01	Pov idle 30% %wt %wt ppm C HAP/HC ppm C HAP/HC 9.7 0.00001 0.76 0.00002 8.9 0.00001 0.64 0.00001 0.2 0.00000 0.012 0.00000 0.2 0.00000 0.034 0.00000 <0.01	Power Setting Idle 30% 75' %wt %wt 75' ppm C HAP/HC ppm C HAP/HC ppm C 9.7 0.00001 0.76 0.00002 0.32 8.9 0.00001 0.64 0.00001 0.27 0.2 0.00000 0.012 0.00000 0.012 0.2 0.00000 0.034 0.00000 0.026 <0.01	Power Setting Idle 30% 75% %wt %wt %wt ppm C HAP/HC ppm C HAP/HC 9.7 0.00001 0.76 0.00002 0.32 0.00009 8.9 0.00001 0.64 0.00001 0.27 0.00000 0.2 0.00000 0.012 0.00000 0.012 0.00001 <0.01	Power Setting Idle 30% 75% 100%(m %wt %wt %wt 9pm C PPm C HAP/HC ppm C HAP/HC ppm C PPM C <td< td=""></td<>	

*C.W. Spicer, M.W. Holdren, S.E. Miller, D.L. Smith, R.N. Smith, D.P. Hughes. "Aircraft Emissions Characterization," Final Report, March 1988, Engineering and Services Laboratory, Air Force Engineering & Services Center, Tyndall Air Force Base, ESL-TR-87-63.

From Table B-5, ATSDR selected the following 6 chemicals (Table B-6) to model based on emission rates and toxicity.

Table B-6. Selected chemicals in TF33-P3 exhaust for air dispersion modeling. One TF33-P3 Engine with JP-4 Fuel

Power Setting								
	ldle %weight(wt)		30% %wt		75% %wt		100% %wt	
	ppm C	HAP/HC	ppm C	HAP/HC	ppm C	HAP/HC	ppm C	HAP/HC"
1,3-Butadiene	11.981	0.00920	0.571	0.00710	0.024	0.00387	<0.001	0.00030
Benzene	12.499	0.00924	1.698	0.02032	0.16	0.02481	0.029	0.00849
Naphthalene	10.138	0.00738	0.395	0.00465	0.035	0.00534	0.05	0.01442
Formaldehyde	15.54	0.02650	4.009	0.11068	0.423	0.15126	0.083	0.05607
Acetaldehyde	1.802	0.00226	1.564	0.03168	0.211	0.05535	0.036	0.01784
Acrolein	1.833	0.00195	0.501	0.00861	0.051	0.01135	<0.001	0.00042
**Non-detects we	ere converted to %	weight base	ed on the a	letection leve	el of 0.001	% weight		

ATSDR performed a screening air dispersion model by selecting a single discharge point on the runway for all emissions. Results of this screening model indicated the need for a more detailed modeling effort. To account for emissions during the movement of aircraft around the base during operations, a more detailed modeling effort was initiated using the following assumptions about where the aircraft were modeled (aircraft mode) and how long they stayed at each location, how long they spent in each engine thrust mode, and the engine settings and corresponding emissions at each location and during each engine thrust mode.

Identifying the Movement of Aircraft.

The location of the jet engines as they operated is important for determining the dispersion of the emissions. These locations would have included the runways, taxiways, parking areas, maintenance areas, approach and takeoff routes, and other areas. The changes of these locations would also be important as well as the routes each type of plane may have used. One important change was the use of two runways. Runways 15/33 and 14/32 were operated together from 1951 through the mid 1960s. These two runways were operated simultaneously. Runway 14/32 was closed in the mid 1960s [45].

Because of the very limited information about aircraft movement, ATSDR simplified the emission locations to Runway 15/33 and the single 10,000 foot taxiway parallel to and just east of Runway 15/33. ATSDR calculated emissions after takeoff and on approach for up to 6 miles. The

aircraft mode and the modeled locations are shown in Table B-7.

Forty-eight volume sources were used to represent taxiway emissions. Fourteen were used to represent takeoffs. Thirty were used to represent climbout. Eighty were used to represent approach. These sources represent aircraft movement at approximately 3-second intervals. Volume sources in each mode (taxi, takeoff etc) were spaced along a line according to their respective speed during that mode.

	Engine		Modeled
Aircraft Mode	Thrust*	Minutes**	Location
Startup	Idle	20	Taxiway
Outbound Taxi	Idle	9	Taxiway
Engine Check	Military for B52	4.5	End of runway
	Military and		where takeoff roll
	afterburn for F16		began
Runway roll	Afterburn for F16	0.7	
	and Military for		Runway
	B52		
Climbout I	Afterburn for F16	0.7	
	and Military for		Straight trajectory
	B52		from runway
Climbout II	Military	0.8	
Approach I	Idle	3	Straight trajectory
Approach II	Idle	1	into runway
Landing on	Idle	1	Dupulou
runway			Kullway
Inbound Taxi	Idle	12	Taxiway
Idle at shutdown	Idle	4.8	Taxiway

Table B-7. Aircraft mode and modeled locations.

* USAF Aircraft Engine Emissions Estimator, Glen D. Seitchek, Air Force Engineering and Services Center, HQ AFESC/RDVS, Tyndall AFB, Florida, November 1985.

**USAF Aircraft Pollution Emission Factors and Landing and Takeoff (LTO) Cycles, Dennis F. Naugle and Steven R. Nelson, Air Force Weapons Laboratory, Kirtland Air Force Base, New Mexico, February 1975.

The direction the aircraft take off and land is dependent on the wind direction. Aircraft take off and land into the wind. For this analysis, bi-directional takeoffs and landings were incorporated. Using climatological data for San Antonio which show the prevailing wind direction to be from the southeast during the summer months (March through September), it was assumed that takeoffs and landings occur from north/northwest to south/southeast during this period [46]. During winter months (October through February), takeoff direction and landings were reversed (south/southeast to north/northwest) since the prevailing wind direction reverses to flows from the north and northeast.

Meteorology

Meteorological data for the ISCST3 model were obtained through U.S. EPA from the San Antonio International Airport for surface data and the Del Rio International Airport, Del Rio Texas for upper air data.

Modeling Process

• All emissions were modeled as volume sources in the ISCST3 dispersion model.

- The volume sources were shifted 30 meters behind the assumed aircraft location to account for jet blast displacement.
- Emissions from all four aircraft modes (taxi, takeoff, climbout, and approach) were included in dispersion modeling.
- The concentrations of the six organic chemicals were estimated at 5,100 points in and around Kelly AFB. The points were distributed 300 meters apart.
- Emissions were calculated based on 336,000 annual operations. Hourly emissions (in g/s) were calculated from an hourly operations value of 19.2 landings and takeoffs per hour. Touch-and-go operations were modeled as a landing and takeoff because the number of touch-and-go operations were not included as a specific number. This means that the 336,000 annual operations were divided into 168,000 takeoffs and 168,000 landings.
- Takeoff and Climbout power settings (and associated fuel flow and hydrocarbon emissions) were set to 100%. Approach and taxi power settings were set to "idle" as described in the previous section. Modeled taxiways were limited to the single 10,000 taxiway parallel to and just east of Runway 15-33.
- Initial horizontal dispersion parameters were assumed 20 meters (estimated at one-third of the B52H wingspan).
- Initial vertical dispersion parameters were assumed 30 meters, based on a review of Photographic Measurements of USAF Aircraft Plume Rise (Music P D, Hunt J S, Naugle DF. Civil and Environmental Engineering Development Office Tyndall AFB FL Detachment 1 [ADTC]] Report Number CEEDO-TR-77-57).

• A release height of 2 meters was assumed for taxiway and takeoff sources.

• After the initial displacement from the engine (30 meters up and down and 20 meters side to side) the plume was considered to be at ambient temperature.

Source release heights for climbout sources varied from approximately 45 to 1375 meters.

- Source locations along the climbout track were calculated along a projected path computed from a 110 knot climbout speed (assumed), a 3000 fpm climbout rate (assumed), and the 1.5 minute time-in-mode (from EDMS)[47].
- Source release heights for approach sources varied from about 480 meters to 0 meters above ground. Source locations along the approach track were calculated along a projected path computed from the 4 minute time-in-mode (from EDMS), a 75 knot approach speed (assumed), and a 3 degree glide slope [48].
- Source locations for runway role were based on accelerating motion. A beginning speed of approximately zero knots and an ending speed of approximately 110 knots was used to calculate source locations along the runway at 3-second intervals. A 42 second time-in-mode (from EDMS) was used.
 - A 46-minute taxi time was used based on data presented in a previous section. The taxi time is the total time for taxi during takeoff and taxi during landing (see Table B-4)and includes time for startup (20 minutes), outbound taxi (9 minutes), inbound taxi (12 minutes), and idle at shutdown (4.8 minutes). This data was obtained from USAF Aircraft Pollution Emission Factors and Landing and Takeoff (LTO), Dennis Naugle, et al, AD/A-006 239 (February 1975).

A constant speed of 2.5 knots was assumed for aircraft movement along the taxiway.

Forty-eight volume sources were used to represent taxiway emissions. Fourteen were used to represent takeoffs. Thirty were used to represent climbout. Eighty were used to represent approach. These sources represent aircraft movement at approximately 3-second intervals. Sources in each category were spaced according to their respective speed during that mode.

Sensitivity Analysis

ATSDR modeled air emissions from a worst-case aircraft (B52) and a best case aircraft (F16) to describe the possible range of air emissions regardless of the specific types of planes that actually were responsible for air emissions. Modeling a fleet of known aircraft would result in emissions that are expected to be within the range estimated by modeling the worst and best cases. Using the revised scenario described herein at Kelly AFB, the range of possible values changed from a factor of 2 to a factor of about 5, with the worst-case values remaining the same.

ATSDR reran the modeling of the F16 emissions with a change in the emission rate during engine check. The engine check emission rate used and evaluated in the Past Air Emissions Health Consultation for Kelly Air Force Base was set at afterburner mode (also called stage 1 augmentation). The emission rate was changed to military power (also called 100% power) in the analysis herein. Engine check emission for the afterburner setting consisted of 3.7 g/s for

benzene and butadiene. Engine check emissions for military power consisted of 1.1 g/s for benzene and 0.92 g/s for butadiene. From the modeling, Figure B-4a and Figure B-5a were redrawn as Figure B-4b and Figure B-5b. The change in this emission rate reduced the predicted concentrations by a factor of 2 for the F16. Table B-8 shows the results for the F16. With the revision, the difference in risk between the F16 and the B52 is about 4 times for butadiene and 5-6 times for benzene.

The U.S. Army Center for Health Promotion and Preventive Medicine Environmental Health Risk Assessment Program (CHPPM), in response to our report, modeled past aircraft emissions using the Federal Aviation Administration's EDMS model[49]. ATSDR has recently been given a draft document. The report includes modeled ambient air concentrations from aircraft emissions but does not include calculations of cancer risk. CHPPM's predicted air concentrations from B52 emissions are within 10% of ATSDR predictions. The B52 was used as a worst case (largest emitter) to determine if further evaluation was necessary. This indicates that the type of model used here is not important in how the results were generated.

The CHPPM also predicted air concentrations from a "more realistic" fleet of aircraft [15] which was not available to ATSDR at the time the work on this report was initiated. CHPPM's results using a fleet of planes were much lower than ATSDR's least emissions scenario using the F16 aircraft. The differences are most likely due to assumptions in the methodology used in creating emission factors for the fleet of planes. ATSDR used a similar methodology for the F16 as the B52 including F16 plane and engine specific emission factors and chemical speciation of the exhaust. CHPPM used extrapolations of the B52 speciation combined with engine specific hydrocarbon emission rates. The emission rates for each plane type are directly proportional to the modeled air concentrations. Therefore, the output is very sensitive to the emission rates that are used. ATSDR will consider the results and methodology of the CHPPM report when it becomes final.

The emission rates are a function of the engine emission rates per engine per time. The number of operations also influences these values. There is some concern that the number of operations used in this report overestimates actual operations. In this evaluation, ATSDR defined an operations as a takeoff or a landing including startup, shutdown, taxing, engine check, runway roll, take off, landing and approach [40]. Operations could include aircraft flying through airspace controlled by Kelly AFB [15] or other movements on the ground. ATSDR's approach could overestimate actual emissions.

Kelly AFB operated a second runway from the 1950s up to the mid 1960s. Modeling the emissions from this runway would reduce the maximum concentrations at the point of maximum exposure. The type of aircraft and number of operations using the second runway are not known.

Given the uncertainties inherent in the analysis, the revision would not result in changes in conclusions and recommendations from a public health perspective.

ATSDR's evaluation of air emissions from jet engine exhaust focused on benzene and 1,3-

butadiene because they contribute the great source of cancer risk. Once emitted from the jet engines these chemicals are transported with the wind and undergo transformation and degradation in the atmosphere.

Benzene and 1,3-butadiene both undergo transformation in the urban air from reactions with hydroxyl radicals (from photochemical reactions), ozone, and other atmospheric chemicals. The half-life of benzene has been measured from 1.5 hours in a "polluted air" to 5 days [19]. 1,3-butadiene is considered more reactive. Half-lives for 1,3-butadiene reported in the literature vary by the type of studies and range from 1.4 to 14.9 hours as shown below attributed to specific atmospheric components [20]:

Half Life (hours)	Description
6	By photochemical produced hydroxyl radicals
2 to 6	Photodegradation
1.4 to 1.7	By ozone (average atmospheric concentration, probably higher concentration in cities-faster degradation)
15 to 16	Triplet Oxygen (average atmospheric concentration)
14.9	Night time degradation from the average atmospheric nitrate radicals concentration

ATSDR evaluated the significance of the degradation by modeling jet emissions of 1,3-butadiene using a 1-hour half-life and a 9-hour half-life. These half lives values were used based on a report by the California Air Resources Board that stated "[a]tmospheric half-lives of 1 to 9 hours are expected." [50]. This range was reasonable to evaluate as 1 hour was near the lower end reported. 9-hours was reasonable to use as a higher value because it is in the range of the higher values. The results show that higher half-lives would not significantly change the concentrations near the base where the population of interest resides because the travel time of air emissions is much faster than 9-hours or 540 minutes (Figure A-1 demonstrated this for hexavalent chromium). The model was run with no degradation as a worst case.

Using an average of the air dispersion modeling results with half-lives of 1 and 9 hours, the general effect is to move the northern edge of the 1E-4 contour line about 0.4 miles closer to the base. The northern edge of this contour is about 1 mile north of the Kelly AFB boundary that lines up with U.S. Interstate 10. With an average half-life (5 hours), the contour line would become about 0.6 miles from the base. This movement or contraction of the risk contour becomes smaller the closer to the emission source which is the runway and taxiway in this case. The changes in the half-live would not result in changes in conclusions and recommendations from a public health perspective.

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Results of air dispersion modeling and conclusions about estimated levels of individual contaminants.

Results

ATSDR performed an air dispersion model to estimate potential levels of contaminants in the community. As addressed earlier, ATSDR does not have information on all of the aircraft that performed takeoffs and landings at Kelly AFB. ATSDR modeled emissions from aircraft for which information on emissions could be located and ATSDR considers to be representative of a range of potential emissions from different aircraft. ATSDR modeled emissions from an F16 aircraft and a B52H aircraft to attempt to approximate the range of potential emissions. An F16 has one engine and a B52H has eight engines. The B52H emissions were considered the worst-case emissions. Benzene, 1,3-butadiene, and formaldehyde were identified as contaminants of concern from worst-case modeling of B52H emissions. The predicted levels of benzene and 1,3-butadiene from air modeling of the emissions from a B52H aircraft are shown in Figures B-2 and B-3, respectively. The predicted levels of 1,3-butadiene from air modeling of the emissions from a B52H aircraft are shown in Figures B-4a and B-4b and for benzene in Figures B-5a and B-5b. (Figures B-4a and B-5a use a afterburner stage 1 emissions factor during engine runup.)

Location of maximum formaldehyde levels are not shown but are located at the same locations as benzene and 1,3-butadiene maximums. All estimated off-base contaminant levels were below levels where health effects have been reported in the scientific literature. Community exposures of modeled annual average concentrations were below levels of concern for acute and chronic noncancer health effects, except for potentially irritating or exacerbating respiratory effects from exposure to formaldehyde. These effects are likely short-term and possibly periodic depending on the changing level of aircraft operations. Benzene, 1,3-butadiene, and formaldehyde were the only chemicals of concern for an estimated increase in cancer risk.

Formaldehyde is produced during combustion of fossil fuels and is also endogenously produced by cellular respiration. While formaldehyde is considered a nasal cavity carcinogen in animals at high doses, evidence in humans is limited. Some epidemiological studies have associated formaldehyde exposure in industrial workers with site specific respiratory cancers while another suggests an increased risk for leukemia [21, 51–53]. Some epidemiological studies also suggest an increased risk for hematopoietic cancers in medical workers and embalmers [22]. Exposure to worst-case conditions during the period of elevated aircraft operations resulted in an increase in the risk for developing cancer for that period of time. Continuous exposure is averaged over a year because data is not available to more discretely define the exposure. It is possible that exposures occurred to higher levels for shorter periods of time, much like an occupational exposure. The Occupational Safety and Health Administration (OSHA) permissible exposure levels (PELs) are presented for perspective, which describe levels at which workers may be exposed for 8 hours per day for 5 days per week (Table B-8).

The cancer risks for 1,3-butadiene reflect a low to moderate increase in risk, depending on the cancer slope factor and the aircraft used as a source of emissions for input in the air modeling. The greatest variable is the use of an appropriate cancer slope factor used to calculate an estimate

of risk. The human-derived cancer slope factor for 1,3-butadiene was based on only one study with human data. Although ATSDR gives preference to human-derived values over animalderived values, the uncertainty in this derivation promotes little confidence in the sole use of this value. ATSDR presented both slope factor values (animal-derived and human-derived) in risk calculations using both F16 and B52H emissions. The resulting risk estimates differed by over 2 orders of magnitude, which illustrates the degree of uncertainty. Table B-8 shows the estimated maximum risk from a continuous off-base exposure to modeled concentrations estimated for 20 years prior to 1973 using maximum operations (336,000/year) and 20 years after and including 1973 (using 112,000 operations/year) to 1994, when JP-8 jet fuel replaced JP-4 jet fuel. Continuous exposure to maximum average concentrations reflects the conservative nature of these estimates. Using a worst-case scenario is likely to overestimate the actual risk.

Benzene risk ranged from low to no apparent increase in risk, depending on the aircraft used as a source of emissions in air modeling. The cancer slope factor for benzene is less uncertain than with 1,3-butadiene as the information for the slope factor was available from several different human studies. ATSDR has developed suggested guidelines which are used to evaluate benzene exposures in air [54]. If the exposure is less than $32 \ \mu g/m^3$, ATSDR assumes there is no apparent public health hazard. If exposures occur between $32 \ \mu g/m^3$ and $320 \ \mu g/m^3$, ATSDR evaluates these on a site-by-site basis. An exposure greater that $320 \ \mu g/m^3$ may be considered a potential health hazard. Although the estimated levels of benzene were below levels at which ATSDR would have public health concerns, the uncertainty in the available data and the elevation of leukemia incidence in ZIP Codes 78227, 78237, and 78226 (1990–1994) indicate that further evaluation is warranted. ATSDR's evaluation on leukemia can be located in the Health Outcome Data Evaluation Health Consultation [25].



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FIGURE B-4a



46

FIGURE B-4b



FIGURE B-5a



48
FIGURE B-5b



Discussion

1,3-butadiene, benzene, and formaldehyde were the chemicals generating the highest cancer risk. These levels of chemicals are below levels where health effects have been reported in workers.

Workers are considered the healthiest subpopulation of the general population. The Occupational Safety and Health Administration (OSHA) and the National Institute for Occupational Safety and Health (NIOSH) regulate the level of contaminant to which workers can be exposed and not be expected to develop adverse health effects (see Table B-8, Worker Exposure Levels). Scientists often do not know at what level more susceptible individuals of the general population might develop health effects, as most information comes from animal studies or epidemiological studies of workers.

Noncancer health effects usually exhibit a threshold effect below which adverse health effects are unlikely. For noncancer health effects, ATSDR develops comparison values (minimum risk levels [MRLs]), which are below threshold levels at which even the more susceptible individuals in the population are more likely to develop health effects. Noncancer health effects have not been reported in the scientific literature at the average annual contaminant levels estimated to have been present in the community around Kelly AFB. Data is not available to evaluate excursions from the annual average. ATSDR used the maximum off-base annual average for evaluation.

Most scientists assume that there may be no threshold for the initiation event in the development of cancer. Most cancers have been studied in groups of people like workers. Scientists predict the probability of developing cancer mostly from these epidemiological studies of workers and from animal laboratory studies. Because cancers often involve long latency periods, it may be 10–30 years before the cancers are diagnosed, so scientists express the risk of developing cancer through a risk assessment. Each individual has a different risk because each individual has different risk factors, including genetics, illness, diet, environmental exposures, occupational exposures, and home exposures. The scientist cannot predict who may or may not develop cancer from an environmental exposure because the scientist cannot know the risk factors for each individual. The risk expressed by the scientist refers to the upper bound risk for an individual in the general population, but the individual's actual risk is unknown. An individual's actual risk may be as low as zero or may be somewhat higher than the estimated risk. Risk assessment is used to describe the *relative* degree of hazard from an exposure, but may not be strictly applicable to the individual that was exposed.

Figures B-6 and B-7 depict levels of interest for benzene and 1,3-butadiene, respectively, and the corresponding risk estimates. Levels of benzene in the community estimated by modeling air emissions are

- in the low risk range,
- similar to the overall national exposure, and
- about 20–30 times less than levels where health effects have been reported.

Levels of 1,3-butadiene in the community estimated by modeling base air emissions are

- in the low to moderate risk range,
- near levels found in a smoke-filled bar,
- less than found around petrochemical plants, and
- about 100 times less than levels where health effects have been reported.

Levels of formaldehyde in the community estimated by modeling base air emissions are

- in the low risk range,
- similar to residential indoor air in conventional homes,
- less than residential indoor air in mobile homes, and
- about 10 times less than levels were health effects have been reported.

Exposures at these levels represent estimates of exposure <u>only</u> to emissions from Kelly AFB. These estimates do not include potential benzene, 1,3-butadiene, and formaldehyde exposure to emissions from other sources, such as automobile and other industrial emissions, and building materials. A person's total exposure may include inhalation of benzene, 1,3-butadiene, and formaldehyde from other sources. ATSDR used the *maximum* off-base annual average concentrations for evaluation.

Figure B-8 is included for comparison and depicts the location and magnitude of the cumulative risk from exposure to *current* (1995 and after) air emissions and *current* (1995 and after) industrial emissions. Current aircraft emissions are an average of the B52 and F16 emissions using JP-8 jet fuel and current level of operations (60,000 operations per year).

Chemical Scenario		Maxin Off-I Concen (µg/	mum Dase tration m ³) ^c	Chronic Non- Cancer Comparis	Cancer Comparison	Worker Exposure	Estimated Cance Risk	
		before 1973	1973 to 1994	on Value (µg/m)	vanue (μg/m³)	μg/m ³)	before 1973	1973 to 1994
1,3-butadiene	F16 human data ^a	10 [4.4] ^j	3 [1.5]	NA			1E-05 [8E-06]	4E-06
1,3-butadiene	F16 animal data ^b	10 [4.4]	3 [1.5]	NA			8E-04 [4E-04]	2E-04
1,3-butadiene	B52 human data	20	7	NA	0.004 °	2200 ^d	3E-05	9E-06
1,3-butadiene	B52 animal data	20	7	NA			2E-03	6E-04
benzene	F16 human data ^f	10 [4.5]	3 [1.5]	13 ^h intermediate	0.14	200.5	2E-05 [2E-05]	7E-06
benzene	B52 human data	20	7	13 ^h intermediate	0.1 *	320 ^s	5E-05	2E-05
formaldehyde	B52	58	19	10 ^h	0.08 ^k	922 ^d	2E-04	7E-05
acetaldehyde	B52	5	2	9 ⁱ	0.5 ^k	360,000 *	3E-06	1E-06
naphthalene, methyl- naphthalenes	B52	16	5	10 ^h	-	50,000 ^g		
acrolein	B52	4.2	1.4	0.02 ⁱ	_	250 ^g		

Table B-8. Estimated maximum past off-base average annual ambient air concentrations from stationary and aircraft emissions.

NA Not Available a Cancer Slope

Cancer Slope Factor (4.3E-6/µg/m³) derived from human data [External Review Draft - Health Risk Assessment of 1,3-Butadiene. US EPA. NCEA-W-0267. January 1998. National Center for Environmental Assessment. Office of Research and Development. Washington, DC.]. All risk estimates assume continuous 20 year exposures before 1973 and continuous 20 year exposures from 1973 to 1994 to the maximum annual average concentrations for each era and 336,000 operations/year before 1973 and 112,000 operations/year from 1973 to 1994.

- b Cancer Slope Factor (0.00028/µg/m³) derived from animal data [IRIS].
- d Occupational Safety and Health Administration Permissible Exposure Level.
- e ATSDR Cancer Risk Evaluation Guide
- f Cancer slope factor 7.8E-06 µg/m³ (EPA IRIS)
- g National Institute for Occupational Safety and Health Time Weighted Average.
- h ATSDR Minimum Risk Level
- i EPA Inhalation Reference Concentration (RfC).
- j. These concentrations are estimates based on using an engine setting of 100% power (military setting) during engine check. The other concentrations and cancer risk estimates are based on an engine setting of afterburner stage 1.
- EPA Cancer slope factors from IRIS. Formaldehyde : 0.000013/µg/m³ from animal data, as no human data is available; Acetaldehyde: 2.2E-06/µg/m³ from animal data, as no human data is available.



Figure B-6 Comparative Levels of Benzene

- a ATSDR CREG (Cancer Risk Evaluation Guide)
- b Wallace, LA. 1989. Major sources of benzene exposure. Environ Health Prospect 82:165–169.
- c Brunnemann KD, Kagan MR, Cox JE, et al. 1989. Determination of benzene, toluene and 1,3-butadiene in cigarette smoke by GC-MSD. Exp Pathol 37:108–113.
- d ATSDR draft guidelines for benzene (Division of Health Assessment and Consultation).



Figure B-7 Comparative Levels of 1,3-Butadiene

ATSDR CREG (Cancer Risk Evaluation Guide)

а

b

Airometric Information Retrieval System. 1994. San Antonio, TX.

c Airometric Information Retrieval System. 1988. Houston, TX.

d Brunnemann KD, Kagan MR, Cox JE, et al. 1990. Analysis of 1,3-butadiene and other selected gas-phase components in cigarette mainstream and sidestream smoke by gas chromatographymass selective detection. Carcinogenesis 11:1863.

e Texas Air Control Board. 1990. Written communication to Bill Henriques (ATSDR), regarding 1,3-butadiene concentrations in air. Austin, Texas.



Appendix B, Attachment 1

Emission rates and locations of past stationary air emissions.

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
Benzene			
BLDG_003	643403.44	4171958.25	3.716963E-04
BLDG_009	642750.25	4171896.25	7.666236E-04
BLDG_036	642130.06	4170109.0	1.215643E-08
BLDG_053	641198.88	4169204.75	4.706397E-05
BLDG_062	641198.88	4169204.75	2.416111E-05
BLDG_063	640863.15	4169672.55	2.013426E-05
BLDG_065	640544.69	4169687.13	4.108452E-03
BLDG_082	639870.56	4170894.0	5.808811E-05
BLDG_086	639651.5	4171084.25	1.013861E-05
BLDG_089	639689.0	4171134.75	1.977116E-06
BLDG_096	642721.25	4171791.25	1.678367E-05
BLDG_098	639679.5	4172361.5	7.544674E-08
BLDG_114	640029.56	4172104.5	1.990896E-05
BLDG_142	641055.81	4172696.5	2.519330E-05
BLDG_159	641769.35	4173519.08	2.643540E-07

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
1,3-Butadiene			×
BLDG_655	640544.69	4169687.13	1.580000E-02

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
Chromium-Sources used in se	nsitivity analysis		
BLDG_030	642041,27	4171147.04	2.241729E-04
BLDG_032	642461.19	4170806.06	1.807650E-04
BLDG_033	642293.13	4170746.0	8.631835E-07
BLDG_036	641652.64	4171415.92	4.574073E-04
BLDG_037	641799.0	4170576.0	2.100413E-04
BLDG_038	643401.38	4171946.25	2.100413E-04
BLDG_053	641198.88	4169204.75	4.791997E-06
BLDG_064	640479.69	4169650.75	2.877278E-05
BLDG_065	640544.69	4169687.13	6.252336E-04
BLDG_082	639870.56	4170894.0	3.740462E-05

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
Chronium-Sources used in s	ensitivity analysis		
BLDG_114	640029.56	4172104.5	1.086371E-05
BLDG_205	638908.31	4170765.0	2.913244E-05

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
Ethylene Benzene		•	
BLDG_003	643403.438	4171958.250	1.0624e-04
BLDG_004	643516.438	4172127.500	3.9676e-05
BLDG_008	643160.375	4171753.250	3.4500e-06
BLDG_009	642750.250	4171896.250	2.1913e-04
BLDG_030	642039.062	4171204.250	1.7250e-06
BLDG_032	642529.188	4170732.000	7.3314e-05
BLDG_035	642120.062	4170665.000	6.9000e-06
BLDG_036	642130.062	4170109.000	1.9200e-07
BLDG_037	641799.000	4170576.000	2.9325e-05
BLDG_038	643401.375	4171946.250	2.9325e-05
BLDG_053	641200.875	4169207.750	1.3700e-06
BLDG_062	640798.750	4169691.750	8.0200e-07
BLDG_063	640866.750	4169669.750	8.0200e-07
BLDG_064	640479.688	4169650.750	1.7250e-06
BLDG_065	640542.688	4169685.750	5.1440e-06
BLDG_082	639870.562	4170894.000	3.7088e-05
BLDG_086	639651.500	4171084.250	2.8979e-06
BLDG_089	639690.500	4171132.250	1.1680e-10
BLDG_094	638908.312	4172108.500	8.6300e-07
BLDG_096	642721.250	4171791.250	4.7970e-06
BLDG_098	639694.500	4172342.500	4.3500e-07
BLDG_114	640029.562	4172104.500	7.0581e-10
BLDG_142	641055.812	4172696.500	7.0581e-10
BLDG_159	641624.000	4173499.750	7.0581e-10
BLDG_205	638908.312	4170765.000	3.4500e-06

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
Methyl Ethyl Ketone			
BLDG_005	643305.375	4171849.250	2.7332e-03
BLDG_008	643160.375	4171753.250	5.1050e-02
BLDG_030	642185.594	4171098.625	8.9946e-02
BLDG_032	642529.188	4170732.000	8.6238e-01

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
Methyl Ethyl Ketone			
BLDG_033	642133.062	4170778.000	6.6980e-04
BLDG_035	642120.062	4170665.000	2.4310e-03
BLDG_036	641786.000	4171156.200	4.7493e+00
BLDG_037	641799.000	4170576.000	1.2422e+00
BLDG_038	643401.375	4171946.250	8.1559e-01
BLDG_052	641148.344	4169358.750	2.1636e-01
BLDG_064	640468.355	4169635.083	4.7404e-02
BLDG_082	639898.562	4170929.333	5.4028e-01
BLDG_090	639360.407	4171818.750	1.0908e-01
BLDG_094	638908.312	4172108.500	1.0028e-02
BLDG_141	642695.250	4171616.250	3.6464e-03
BLDG_142	641066.812	4172716.500	1.8232e-02
BLDG_205	638908.312	4170765.000	1.9752e-02

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
Methylene Chloride			
BLDG_032	642529.188	4170732.000	0.810127767
BLDG_033	642268.125	4170467.000	0.012274663
BLDG_035	642120.062	4170665.000	0.024549326
BLDG_036	641740.000	4171439.250	2.172615401
BLDG_037	641810.500	4170497.500	69.750773318
BLDG_038	643401.375	4171946.250	1.687766181
BLDG_062	640817.750	4169672.667	2.780635652
BLDG_063	640863.150	4169672.550	2.317196381
BLDG_064	640468.355	4169635.083	4.05063884
BLDG_082	639870.562	4170894.000.	0.699655799
BLDG_142	641066.812	4172716.500	0.024549326
BLDG_005	643305.375	4171849.250	0.009974559
BLDG_030	642035.062	4171149.750	10.2891548
BLDG_032	642516.188	4170779.000	12.24327768
BLDG_036	641616.438	4171349.250	8.955193021
BLDG_037	641850.000	4170618.000	8.583610039
BLDG_062	640817.750	4169672.667	1.537964586
BLDG_063	640863.150	4169672.550	1.281637078
		; 	
Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)
Tetrachloroethylene (PCE)			
BLDG_005	643305.375	4171849.250	0.009974559
BLDG 030	642035.062	4171149.750	10.2891548

•

Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)			
Tetrachioroethylene (PCE)	Tetrachioroethylene (PCE)					
BLDG_032	642516.188	4170779.000	12.24327768			
BLDG_036	641616.438	4171349.250	8.955193021			
BLDG_037	641850.000	4170618.000	8.583610039			
BLDG_062	640817.750	4169672.667	1.537964586			
BLDG_063	640863.150	4169672.550	1.281637078			
Building Number	Easting (meters)	Northing (meters)	Emission Rate (g/s)			
Toluene						
BLDG_003	643403.438	4171958.250	0.0027391			
BLDG_004	643516.438	4172127.500	0.0022311			
BLDG_005	643305.375	4171849.250	0.00022168			
BLDG_008	643160.375	4171753.250	0.00089195			
BLDG_009	642750.250	4171896.250	0.0056494			
BLDG_030	642185.594	4171098.625	0.003671			
BLDG_032	642469.167	4170769.000	0.0082629			
BLDG_033	642268.125	4170467.000	0.00055778			
BLDG_035	642120.062	4170665.000	0.0017771			
BLDG_036	641889.825	4170925.950	0.011377			
BLDG_037	641799.000	4170576.000	0.011298			
BLDG_038	643401.375	4171946.250	0.011117			
BLDG_053	641198.875	4169204.750	6.2955E-5			
BLDG_062	640817.750	4169672.667	7.2377E-5			
BLDG_063	640863.150	4169672.550	6.0314E-5			
BLDG_064	640468.355	4169635.083	0.005487			
BLDG_065	640544.688	4169687.125	0.0054956			
BLDG_082	639870.562	4170894.000	0.0058632			
BLDG_086	639651.500	4171084.250	7.4713E-5			
BLDG_089	639689.000	4171134.750	1.457E-5			
BLDG_094	638908.312	4172108.500	0.0001816			
BLDG_096	642721.250	4171791.250	0.00012368			
BLDG_098	639679.500	4172361.500	4.2102E-6			
BLDG_114	640029.562	4172104.500	2.6434E-5			
BLDG_141	642695.250	4171616.250	2.5943E-5			
BLDG_142	641061.312	4172706.500	0.00091206			
BLDG_159	641769.354	4173519.083	6.3239E-6			
BLDG_205	638908.312	4170765.000	0.00048643			

* Final Report. Historical Air Emissions Estimate. Kelly AFB, TX. EARTH TECH, Inc. San Antonio, TX. March 27, 2000.

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Appendix B, Attachment 2



DEPARTMENT OF THE AIR FORCE AIR FORCE INSTITUTE FOR ENVIRONMENT, SAFETY AND OCCUPATIONAL HEALTH RISK ANALYSIS (AFMC) BROOKS AIR FORCE BASE TEXAS

13 February 2001

MEMORANDUM FOR AFBCA/DK

ATTENTION: MR CHARLES WILLIAMS

FROM: AFIERA/RSRE 2513 Kennedy Circle Brooks AFB, TX 78235-5123

SUBJECT: Consultative Letter, IERA-RS-BR-CL-2001-0011, Availability of Information Related to Unburned Fuel and Oil Misting Emissions From Aircraft Takeoffs and Landings

1. The Agency for Toxic Substances and Disease Registry (ATSDR) requested available information related to emissions from aircraft takeoffs and landings. This information is needed to address community and ATSDR public health concerns related to aircraft fuel and oil misting. The specific question posed by ATSDR was whether any available data exists on sampled or modeled unburned fuel or nil mist emissions during takeoffs and landings.

2. We contacted the Federal Aviation Administration (FAA) and several organizations within the US Air Force (USAF): Robert Holsclaw, FAA, Environmental Section, Maj Jeanette Howard, Air Quality Branch (AFIERA/RSEQ), Maj Brian Blazicko (AFIERA/RSHI), and Carry Embree, Propulsion Environmental Working Group Member (OC-ALC/LR). No information was available from these groups related to quantitative analyses of aircraft fuel or oil misting (Wade, 2001). Three recently accomplished studies qualitatively describe fuel and oil misting, as referenced in four documents: Massport International (1997), Hof hoghe et al.(1997), KM Chng Environmental Inc.(1999), (Attachments 1-4). These studies were initiated in response to assertions from residents and government officials that aircraft using O'Hare and Logan airports caused the deposition of soot, particles or oily film on surfaces in communities near these airports. Hoffnagle et al.(1997), KM Chng (1997; 1999), and Massport International (1997) conclude that deposited particles monitored near the airports bore little resemblance to either unburned jet fuel or soot from jet exhaust.

3. ATSDR specifically asked us to obtain information related to unburned JP-8 aerosol exposure to ground erews during cold-engine start conditions. Maj Blazicko (2001) provided the following information regarding these evaluations:

a. The AF Institute for Environment, Safety, and Occupational Health Risk Analysis/Industrial Hygiene Branch's (AFIERA/RSHI) initial testing determined that existing sampling methods were inadequate to properly characterize the exposure due to the volatile nature of the JP-8 aerosol.

b. A current AFIERA/RSHI research effort with the University of North Carolina (Demonstration of Sampling Method for JP8 Aerosols; Contract F41622-97-C-D025) was established to develop and employ a methodology to evaluate ambient JP8 concentrations of aircraft emissions at temperatures of 0°F and below. The UNC study will be completed in the last quarter of 2001. This information is designed for

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Appendix B, Attachment 2

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incorporation into an occupational health risk assessment of ground crews and includes no information on takeoffs and landings.

4. Joe Franzello, Brooks AFB Technical Library, assisted us in querying several databases to determine the availability of information related to emissions of unburned fuel and oil misting from aircrafts. The following databases and keywords were used:

Database	Keywords	
FirstSearch	aircraft and oil mist	
	aircraft and unburned fuel	
	aircraft and soot and exhaust	
	aircraft and carbon black	
DialogTech	Acrosol or atomic or spray or	
	mist and fuel or JP	

No quantitative information was located as a result of these searches (Attachment 5).

5. Evaluations of ground crew exposures may not indicate unburned fuel or oil misting levels, but may assist ATSDR in evaluating their emission models due to the proximity of the ground crews to aircraft taxing, takeoffs, and landings. Ground crew exposures to F-15, KC-135, and C-130 aircraft engine exhaust were evaluated through three studies (Johnston and Fritts, 1999; Johnston and Fritts, 1999b; and Johnston and Lazenby, 1999, Attachments 6-8). Personal samples were collected from ground crew chiefs who are responsible for preparing aircraft for launch and, upon recovery, performing post-flight check procedures. Analyses included aldehydes, BTEXs (benzene, tolkene, ethyl benzene, and xylenes), gases (earbon monoxide, nitric oxide, nitrogen dioxide, sulfer dioxide), JP-8 fuel (as naphthas), and polymuclear aromatic hydrocarbons (particulate and vapor fraction). Several aldehydes analyses were quantifiable. All other results were below their respective detection limits.

6. If you have any comments or questions, please call Mr Jody Wireman at (210)536-6123 or Mr Cornell Long at (210)536-6121,

Lighton.

JODY Ř. WIREMÁN Environmental Scientist Environmental Sciences Branch

Attachments:

- 1. Massport International, 1997 (2 copies)
- 2. Hoffnagle et al., 1997 (2 copies)
- 3. KM Chng Environmental Inc., 1997 (2 copies)
- 4. KM Chng Environmental Inc., 1999 (2 copies)
- 5. Literature Search Results (2 copies)
- 6. Johnston and Fritts, 1999a (2 copies)
- Johnston and Fritts, 1999b (2 copies)
 Johnston and Lazenby, 1999 (2 copies)
- 9. References (2 copies)

Attachment 9 - References

Blazicko, 2001. Blazicko B. Personal communications about JP-8 cold engine startup occupational exposure research project. January 31, 2001.

DialogTech, 2001. DialogTech. http://www.dialogselect.com/tech/ January 22, 2001.

FirstSearch, 2001. FirstSearch. http://www.oelc.org/firstsearch/ January 23, 2001.

Hoffnagle et al., 1997. Hoffnagle GF, Cooper JA, and Morris S. Soot Deposition Study: Logan Airport and Surrounding Communities. TRC Environmental Corporation. January 1997.

Johnston and Fritts, 1999. Johnston DS and Fritts GD. Consultative Letter, IERA-RS-BR-CL-1999-0008, Engine Exhaust Exposure of F-15 Crew Chiefs, Otis ANGB, MA. February 12, 1999.

Johnston and Fritts, 1999. Johnston DS and Fritts GD. Consultative Letter, IERA-RS-BR-CL-1999-0007, Engine Exhaust Exposure of KC-135 Crew Chiefs, Pease ANGB, NH. February 12, 1999.

Johnston and Lazenby, 1999. Johnston DS and Lazenby M. Consultative Letter, AL-OE-BR-CL-1999-0011, Engine Exhaust Exposure of KC-135 Crew Chiefs, Otis ANGB, MA. February 12, 1999.

KM Chng Environmental Inc., 1997. KM Chng Environmental Inc. Soot Deposition Study: Logan Airport and Surrounding Communities. January 1997.

KM Ching Environmental Inc., 1999. KM Ching Environmental Inc. Findings Regarding Source Contributions to Soot Deposition: O'Hare International Airport and Surrounding Communities. December 1999.

Massport International, 1997. Massport International. Summary of Two Logan Soot Studies. January 1997.

Wade, 2001. Wade M. Personal communications about aircraft oil and fuel misting. AFIERA/RSEQ (Karta Technologies). January 19, 2001.

Appendix C Chemical Mixtures Exposure

Appendix C Chemical Mixtures Exposure

General comments about chemical mixtures

Environmental chemical research has mainly centered on toxicity testing and mechanistic studies of single chemicals. This research lead to a better understanding of the interactions of exposure and susceptibility. However, ATSDR recognizes that humans are often exposed to multiple chemicals. Knowledge based on individual chemical exposure and toxicity is often a limiting factor in the human health assessment process. While interactions among some chemicals have been demonstrated at high concentrations, interactions at low environmental levels have not been scientifically demonstrated. Predicting whether chemicals will act in a potentiating, additive, synergistic, antagonistic, or independent manner at environmental concentrations or in the workplace has limitations.

Chemicals mixtures are found in the air we breathe, the food we eat, and the water we drink. With over 80,000 existing chemicals and 2,000 more being added each year, people are exposed to thousands of chemicals in different combinations every day in the home, the environment, and the workplace. Some of these chemicals have similar mechanisms of action or affect the same organ or tissue, so interactions between these chemicals are possible.

Chemical mixtures may contain two or three chemicals of a similar class or more complex mixtures may contain hundreds of chemicals from different classes. These chemicals may express different levels of toxicity and different modes of action. Changes in one chemical caused by another may alter the resultant toxicity from predicted values. Though changes in toxicity have sometimes been described for simple mixtures, understanding the interactions of complex mixtures has not been achieved.

Individual testing of the endless number of potential combinations is virtually impossible. Even if cost were not considered, the number of animals required to perform statistically relevant toxicity tests with multiple doses for multiple exposure periods would be prohibitive. An experiment investigating three chemicals at five different dose levels at only one time point after exposure, would require 125 treatment groups and 750 animals, if only six animals are included in each treatment group. Therefore, it is unlikely that questions concerning chemical mixtures will be answered through traditional animal research in the near future.

Interactions between chemicals can be potentiating, additive, synergistic, antagonistic; or there may be no interaction and thus, independent. ATSDR evaluates the potential for chemical mixtures on a site-by-site basis. ATSDR assumes that chemicals act independently if they have different modes of action, but additively if the modes of action are the same or effects are on the same target organ, unless there is evidence of interaction between the chemicals. For non-cancer effects, ATSDR assumes that a threshold exists for health effects. For cancer effects, ATSDR assumes that a threshold may not exist for genotoxic chemicals.

ATSDR's approach for the assessment of exposure to chemical mixtures included

- (1) Evaluating cumulative exposures by summing the individual risks for each carcinogen in the absence of compelling evidence supporting a greater than or less than additive model. This method of addressing cumulative risks has been externally peer-reviewed and found to be appropriate and relevant [55]. Under this response addition model, the predicted response to the mixture would be simply additive. This model assumes the contaminants act independently. For past exposures to the maximum level estimated by ATSDR air dispersion modeling, the cumulative risk from individual chemical exposures is considered to be a *low* increase in the risk for developing cancer. Almost all of the estimated risk was due to 1,3-butadiene, benzene, and formaldehyde exposure. A cumulative risk of 4E-04 was estimated by summing individual risks (Table C-1). Actual risks would likely be considerably less than this estimate due to the conservative nature of the assessment using a worst-case emissions scenario and continuous lifetime exposure to maximum average annual concentrations, and assuming additive toxicity.
- (2) Evaluating the evidence for potential interactions among the contaminants.

ATSDR investigated several approaches to evaluate interaction and concluded that scientific information was insufficient to compare the chemical mixture as a whole mixture, a similar mixture, or a component mixture. Epidemiological evidence of interaction involving 1,3-butadiene and benzene is inconclusive but recent evidence suggests independent action [56]. Evidence of interaction between formaldehyde and benzene or 1,3-butadiene was not located and formaldehyde appears to exert toxic effects by a different mode-of-action.

In assessing air emissions for potential interactions of chemical mixtures, ATSDR considered potential effects from the co-exposure of benzene and 1,3-butadiene. They were selected because these chemicals:

- represented the greatest estimated risk, considering quantity and toxicity,
- included the same organ system (bone marrow) as a target for carcinogenic effects, and
- epidemiological investigations of workers have reported confounding exposures to chemical mixtures.

Potential interactions between benzene and 1,3-butadiene have not been studied. Occupational exposure to high levels of benzene or 1,3-butadiene have been associated with the development of leukemia [57–59]. The metabolism of benzene and 1,3-butadiene appears to be similar in laboratory animals, with both chemicals metabolized primarily in the liver by the P450 family of enzymes (principally by the P450 isozyme 2E1 at these concentrations) [60–63]. Like benzene, 1,3-butadiene is metabolized to reactive metabolites but the precise mechanism is unknown [64]. Evidence indicates that the same metabolites detected in laboratory animals will be formed in humans, although the rates may be different [65]. Which metabolite(s) is responsible for the

causation of cancer is still uncertain. Differences in measured concentration levels in mice and rats do not explain the differences in cancer in these species. All three metabolites are mutagenic *in vivo* and *in vitro*. Based on the overall evidence from human, animal, and mutagenicity investigations, EPA concludes 1,3-butadiene to be a known human carcinogen [29].

Benzene is a known human carcinogen in humans while 1,3-butadiene shows clear evidence of carcinogenicity in animals and more recent evidence suggests stronger carcinogenic potential in humans [56]. While occupational exposure to high concentrations of benzene is known to increase the risk for developing non-lymphocytic leukemia, high doses of 1,3-butadiene have been associated with cancers at multiple sites in laboratory animals, including hematopoietic cancers such as lymphocytic leukemia [66]. Epidemiological studies suggest that co-exposure to 1,3-butadiene, styrene, and benzene may be associated with leukemia whereas exposure to 1,3-butadiene alone may be associated with lymphosarcomas [29]. Evidence of an association with high doses of 1,3-butadiene and leukemia in occupational studies is often confounded by co-exposure to other chemicals. The strongest evidence for the carcinogenicity of 1,3-butadiene in humans has occurred during co-exposure to styrene and benzene [56].

Occupational studies are evaluated for the relevance of the effect and the chemical mixture. Relevance is evaluated by assessing the temporality, strength of association, consistency, specificity, dose response, and biological plausibility. The recent University of Alabama-Birmingham (UAB) study was found to be particularly relevant to exposures of 1,3-butadiene, styrene, and benzene [56]. The UAB study investigated styrene and benzene exposures as well as 1,3-butadiene and concluded that the observed associations of leukemia with 1,3-butadiene exposure were not due to confounding exposures to the other chemicals. The authors conclude that exposures to 1,3-butadiene alone were associated with leukemia mortality. The dose-response analysis generated by the authors was used by ATSDR to compare to potential exposures around Kelly AFB. ATSDR compared the highest potential exposure period at Kelly AFB (before 1970) to the dose response of the UAB study.

The highest potential exposure period to benzene and 1,3-butadiene at Kelly would have occurred before 1970 based on operational information, type of jet fuel use, and air dispersion modeling of estimated emissions. Levels of 1,3-butadiene in the community are estimated to have been 20 μ g/m³ (9 parts per billion [ppb]). The majority of housing in the communities was started in the 1950s which would equate to a maximum cumulative dosage of 180 ppb-years, assuming a 20-year exposure (1950–1970). Exposures after that time would be much less compared to the time period before 1970 because operations were significantly less (82,000 takeoff and landings/year compared to 330,000/year. See Appendix B). Exposures after 1970 resulted in a cumulative exposure dose of 54 ppb-years for the period 1970–1994. Kelly AFB changed from JP-4 jet fuel to JP-8 jet fuel in 1994. JP-4 jet fuel contained at least 100 times the benzene concentration measured in JP-8 jet fuel [3].

In the UAB cohort, the median cumulative exposure to 1,3-butadiene, styrene, and benzene was 11,200, 7,400, and 2,900 ppb-years, respectively. Among those dying of leukemia, the median cumulative exposure to 1,3-butadiene was 36,400 ppb-years, 200 times greater than the maximum

estimated annual average exposure at Kelly AFB (180 ppb-years). The UAB cohort consisted of workers, generally considered the healthiest segment of the general population.

Benzene is a known human carcinogen with the bone marrow as the primary target organ. Exposures to high concentrations of benzene have been associated with the development of leukemia, primarily acute non-lymphocytic leukemia (ANLL). Levels above 40 ppm-years are considered to increase the risk for developing leukemia in occupational exposures [27]. Occupational exposure (8 hours/day, 5 days/week, 50 weeks/year) to benzene at 40 ppm-years would be mathematically equivalent to a lifetime environmental exposure (76 years) of 120 ppb ($384 \mu g/m^3$). The estimated community exposure to past levels of benzene of 20 $\mu g/m^3$ for 20 years is equivalent to a lifetime exposure to 6 $\mu g/m^3$, or 64 times less than the lowest level of concern reported in epidemiological studies of occupational exposures. The estimated community exposure was also five times less than the level ATSDR considers as no apparent health hazard ($32 \mu g/m^3$) [54].

Occupational studies have reported an association with benzene or 1,3-butadiene exposure and leukemia mortality in workers. Exposure to levels of benzene and 1,3-butadiene estimated with limited data and air dispersion modeling to have been present in the community prior to 1970 would not be expected to result in leukemia mortality in healthy individuals. Susceptible members of the community may be at greater risk for developing hematopoietic perturbations than workers because

- These occupational studies were performed on workers with daily 8-hour exposures who died of leukemia. The potential health effects these same levels might have to more susceptible members of the general population, continually exposed to lower concentrations, is uncertain.
- These occupational studies do not identify the types of exposures which may have resulted in developing disease, as some individuals may be exposed to higher concentrations for shorter periods of time than others.
- These occupational studies reported mortality (death) from leukemia. It is not known if other workers developed disease or incurred reduced quality of life as a result of exposure.
- Scientific studies have not been performed on potential health effects from exposure to a chemical mixture of 1,3-butadiene and benzene.

Although increased risks for leukemia have been found in medical workers and other professionals exposed to formaldehyde, studies in industrial workers, who are thought to have higher exposures, have shown inconsistent associations [21, 22]. Some scientists have concluded that there is little likelihood that formaldehyde can induce toxicity at sites remote from the respiratory tract [67].

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Summary

Estimated levels of past air exposures to benzene, 1,3-butadiene, and formaldehyde:

- are above some comparison values which are levels that ATSDR considers "safe," even to more sensitive populations. Exceeding a comparison value does *not* indicate that health effects would be likely, but indicates additional investigation may be warranted.
- are below levels associated with worker mortality from leukemia.

Cumulative estimated risks for past air exposures to benzene, 1,3-butadiene, and formaldehyde:

- are based on the addition of estimates of individual contaminant risks as interactions have not been demonstrated [56].
- results in an estimated low increase in the risk for developing cancer.

Epidemiological evidence indicates that:

- exposure to high levels of either benzene or 1,3-butadiene is associated with leukemia mortality in workers, but at levels much higher than estimates of past exposures of either contaminant at Kelly AFB.
- exposure to high levels of formaldehyde has been associated with leukemia in medical professionals and embalmers, but results of exposure in industrial workers has not been consistent.

ATSDR concludes that additional investigation is warranted because:

1. The community has been exposed to chemicals which have been associated with cancer in workers.

2. Confidence in the representativeness and comprehensiveness of the data is very low because most of the sampling and analytical data provided by Kelly AFB were collected before regulatory agencies began reviewing data. Exposure scenarios are also uncertain.

3. Health outcome data indicate that a biologically plausible health outcome, leukemia, was elevated (statistically significant) between 1990–1994 in three ZIP Codes, two of which were downwind and the third was off-base military housing.

4. Potential cumulative effects of chemical mixtures like 1,3-butadiene and benzene are unknown.

Chemical	Scenario	Estimated Cancer Risk⁴	
1,3-butadiene	B52 human data ^a	4E-5	
benzene	B52 human data ^b	7E-05	
formaldehyde	B52 animal data °	3E-04	
Total Estimated Cumulative Risk 4E-04			

Table C-1. Estimated cumulative cancer risks from benzene, 1,3-butadiene, and formaldehyde exposure to stationary and aircraft historical air emissions.

a Cancer Slope Factor (4.3E-6/µg/m³) derived from human data [External Review Draft - Health Risk Assessment of 1,3-Butadiene. US EPA. NCEA-W-0267. January 1998. National Center for Environmental Assessment. Office of Research and Development. Washington, DC.]

b Cancer Slope Factor $(7.8E-06/\mu g/m^3)$ derived from human data [IRIS].

c Cancer Slope Factor $(0.000013/\mu g/m^3)$ from animal data [IRIS]. No human data available.

d

All risk estimates are based on continuous 20 year exposures before 1973 and from 1973 to 1994 to estimated maximum annual average concentrations during each era. Level of operations was assumed to be 336,000 before 1973 and 112,000 1973 to 1994. Risks were summed for both eras.

Appendix D Response to Comments from External Peer Review

Response to Comments from External Peer Review

1. Does the public health consultation adequately describe the nature and extent of contamination?

<u>Reviewer 1</u>

Comment: The consultation describes the contamination about as well as can be expected. Of course, we would all wish for better data. However, given the overall circumstances, the consultation does an excellent job.

Response: No response needed.

<u>Reviewer 2</u>

Comment: In Background, the authors explained the importance of past air emissions and considered contaminations in this consultation. Also, they described contaminants from industrial and aircraft emissions in pages 9 and 10. The characteristics, emissions, and known adverse health effects of those contaminants are explained in detail in Appendix B. This document adequately described the nature and extent of possible past contamination by activities at Kelly AFB.

Response: No response needed.

Reviewer 3

Comment: More description in the text including criteria or standard levels is needed. Measured data results would also be desirable. Discussion of impacts could be expanded. *Response:* The target audience for this document is the community. So as not to detract from the intended message to be communicated in the text, ATSDR puts as much of the detail as possible in appendixes.

2. Does the public health consultation adequately describe the existence of potential pathways of human exposure?

<u>Reviewer 1</u>

Comment: As with the previous question, the pathways of human exposure are fraught with uncertainty in this particular assessment. However, the consultation does an excellent job of describing the major concerns for the potential pathways. *Response:* No response needed.

Reviewer 2

Comment: The authors described the potential exposure pathway elements in Table 1 in which only direct airborne exposures are described. Because toxic chemicals and organic carbons (both gaseous and particulate) can adhere to airborne particles and accumulate on them, the deposited and resuspended particles can act as an airborne vector for these compounds. This represents a potential exposure to residents downwind. Even though this indirect exposure seems to be beyond the scope of this consultation, it is worth noting it in the document.

Response: ATSDR agrees that deposition and resuspension may represent a pathway of concern. ATSDR chose to conservatively address the concern attributed to the potential

pathway by assuming the inhalation exposure represented 100% of the exposure from air emissions, not reducing the exposure to account for deposition. Direct 100% inhalation exposure would be a greater exposure than a portion of the exposure from inhalation and a portion of exposure from chemicals deposited.

Reviewer 3

Comment: Discussion of potential pathways could be expanded. Sources are defined well, but exposure is not directly addressed.

Response: This document is one of several documents prepared to assess environmental exposures at Kelly AFB. The potential pathways were discussed in the Phase I document. This document focused on past air emissions.

3. Are all relevant environmental and toxicological data (i.e., hazard identification, exposure assessment) being appropriately used?

<u>Reviewer 1</u>

Comment: I am not aware of all the potential sources of environmental data for this particular consultation. However, it appears to be a reasonable collection of data (albeit from a highly uncertain history), and the data appear to support reasonable conclusions. More specifically, when conservative assumptions about the data lead to the conclusion of "no significant risk," this is an appropriate use of data. This is a classic screening approach that is "good enough" to answer most of the questions being posed. **Response:** No response needed.

Reviewer 2

Comment: The authors utilized industrial emission data, specific aircraft emission data, and detailed aircraft operational information supplied by Kelly AFB (Appendix B). The hexavalent chromium emission data, jet fuel misting information, and incineration emission data were not available. The model-estimated concentrations were compared with ATSDR chronic non-cancer data, worker exposure data, ATSDR cancer comparison data, and estimated cancer risk data. In pages 7 and 8, the authors clearly explained their selection of contaminants. All environmental and toxicological data were appropriately used for the consultation.

Response: No response needed.

<u>Reviewer 3</u>

Comment: Yes, this has been done but detail could be added. *Response:* The target audience for this document is the community. As such the level of detail is deemed appropriate.

4. Does the public health consultation accurately and clearly communicate the health threat posed by the site?

<u>Reviewer 1</u>

Comment: The consultation does a good job in this area. However, the ATSDR should be prepared to answer the following questions.

Have other groups made estimates of the risks at Kelly Air Force Base? If so, how do they compare with this report?

Are there estimates of the risks at other Air Force Bases? If so, how do they compare with this report?

Response: Kelly AFB through the RCRA and Superfund programs have completed risk assessments on individual operable units and solid waste management units. These risk assessments, however, do not cover past air emissions.

The U.S. Army Center for Health Promotion and Preventive Medicine Environmental Health Risk Assessment Program (CHPPM), in response to our report, modeled past aircraft emissions using the Federal Aviation Administration's EDMS model [49]. ATSDR has recently been given a draft document. The report includes modeled ambient air concentrations from aircraft emissions but does not include calculations of cancer risk. CHPPM's predicted air concentrations from B52 emissions are within 10% of ATSDR predictions. The B52 was used as a worst case (largest emitter) to determine if further evaluation was necessary.

The CHPPM also predicted air concentrations from a "more realistic" fleet of aircraft which was not available to ATSDR at the time the work on this report was initiated. The results of this scenario and its assumptions need to be evaluated further.

ATSDR will consider the results of the CHPPM report when it becomes final.

ATSDR does not know of any other risk estimates from past air emissions from other Air Force bases.

Reviewer 2

Comment: The reviewer cannot judge the accuracy, but the authors clearly described and considered possible health threat posed by Kelly AFB in Appendix C. Also, they reported information on potentially susceptible populations (page 13). *Response:* No response needed.

Reviewer 3

Comment: The threat is accurately described, but more text is required to clearly communicate the threat. Detail should not only appear in tables, but should be discussed. *Response:* ATSDR also provides fact sheets with presentations or availability sessions to clearly communicate messages and provide health education.

5. Are the conclusions and recommendations appropriate in view of the site's condition as described in the public health consultation?

Reviewer 1

Comment: Yes, overall they seem reasonable. However, see question #7 for further comments in this area.

Response: No response needed.

<u>Reviewer 2</u>

Comment: The authors' conclusions for the individual contaminations from stationary sources and aircraft sources are appropriate. Their recommendations of further studies on elevated leukemia outcomes and on-base exposure are relevant. *Response:* No response needed.

<u>Reviewer 3</u>

Comment: Yes, I agree with the recommendations and conclusions. *Response:* No response needed.

6. Given the available information, are the methods used suitable for determining the range of historic ambient air concentrations from aircraft emissions?

Reviewer 1

Comment: They are suitable given the fundamental uncertainty of the data. However, it may be useful for the reader to see a summary of all the assumptions that make ISCST a conservative model for this consultation. The assumptions and conservative nature of this model will be important in interpreting the results.

Response: The assumptions and methodology of the ISC modeling are presented in the discussion of Aircraft Emissions which includes 7 to 8 pages of text. A short summary of these assumptions were added as Table A-1.

Reviewer 2

Comment: All available data and operation information were utilized to estimate downwind concentrations in this consultation. The available air transport models were suitable used. While the authors mentioned ISCST model for the estimation of aircraft emissions, they did not clarify which model they used for the stationary emissions. From the reference (Rodgers et al., J.Exp.Anal.Env. Epi. 9, 535, 1999), the reviewer assumes simple Gaussian dispersion model was used. These steady state plume models (Gaussian model and ISCST model) make an important simplifying assumption, namely that there is no vertical variation in either the wind speed or turbulence intensity. They only consider the standard deviation of Gaussian distribution as dispersion parameters. This drawback often results in overestimations of gaseous or particulate pollutant concentrations (Winges KD USEPA/910/988/202/R, 1990; Kim, E. and Larson, T.V. Atmospheric Environment 35, 3509–3519). It can be one of the possible reason for high estimated concentrations in Tables B-1 and B-8. It looks to the reviewer that Figures B2 and B3 are identical, and Figures B4 and B5 are identical, even though the emissions are different between Butadiene and Benzene. It needs some clarification.

Response: The ISCST3 model was also used for the stationary sources. Appendix B was modified to include this information. The ISCST3 model uses a Gaussian distribution and the dispersion parameters of Pasquill-Gifford. The ISC model includes a variation of wind velocity (EPA 1995; EPA-454/B-95-003a available at

<u>http://www.epa.gov/scram001/tt22.htm#isc</u>) with height. The Fugitive Dust Model (FDM) described by Winges (1990) does not. As the reviewer points out, neither ISC nor the FDM vary settling velocity of particulates with height. For the modeling, ATSDR assumed all emissions were gaseous. ATSDR acknowledges the uncertainty of this assumption but it

presents a worst-case or highest exposure concentration scenario for the modeling of metals and organics.

ATSDR also acknowledges the inherent uncertainty in the ISCST3 model. Using a Gaussianplume model, Rodgers et al. (J.Exp.Anal.Env. Epi. 9, 535, 1999), identify an uncertainty of approximately a factor of 2 in flat terrain (i.e., modeled concentrations range from ½ to 2 times the actual values). This range of uncertainty is relevant in this case because the terrain at Kelly AFB is flat. The Section titled "How did ATSDR evaluate past emissions at Kelly AFB" and Appendix B were modified to include this description of uncertainty.

We are aware that Figures B2 and B3 (predicted air concentrations of butadiene and benzene, from B52 emissions) and Figures B4 and B5 (predicted air concentrations of butadiene and benzene from F16 emissions) appear the same. This occurred because the emission rates of 1,3-butadiene and benzene are similar for the significant mode of operation. For the B52, modes contributing the predominant risk are startup, shutdown, and taxi. The engine setting during these modes is idle. The total emissions for the time the plane is using an idle engine setting are 2544 g/plane benzene and 2534 g/plane butadiene.

For the F16, the mode contributing the predominant risk is engine check and takeoff as seen by the higher concentrations at the ends of the runway. The engine check emissions are 204 g/plane benzene and 173 g/plane butadiene. The takeoff emissions are 305 g/plane benzene and 322 g/plane butadiene.

<u>Reviewer 3</u>

Comment: It would appear so, but more detail and better organization of presented data is needed.

Response: The target audience for this document is the community. As such the level of detail is deemed appropriate.

7. ATSDR identifies a range of risk estimates (identified in Table B-8) for potential past exposures to benzene and butadiene by utilizing the B52 aircraft as a worst-case and the F16 aircraft as a best case emissions scenario. From a public health perspective and considering the uncertainty, would modeling each individual aircraft emissions (more than 50 different types of aircraft) change the conclusions and recommendations?

<u>Reviewer 1</u>

Comment: There is no need to consider further data when the worst-case scenario leads to the conclusion that a public health concern is "unlikely." Lower exposures would yield the same conclusion, and the assumptions about the best and worst case seem very reasonable. This is screening assessment at its best.

The more provocative conclusion -- #3 on page 14 – is that the exposures "may have resulted in an increased risk for developing cancer." Given the uncertainties throughout this consultation (not just with the ISCST modeling), it may be more accurate to say that "the data are inconclusive with regard to cancer risks and follow-up is needed." Both statements may be true, but it is a strategic decision on which statement to use. The former statement may evoke needless fear from the public. In my view, the later statement cannot be seen as overly optimistic, particularly since it calls for follow-up. However, the investigators who are closest to the community will make the most appropriate judgment on this issue.

In either case, follow-up with the Health Outcome Data Evaluation Health Consultation is still very appropriate.

Response: No response needed.

<u>Reviewer 2</u>

Comment: It does not seem to the reviewer that the conclusion and recommendations would change by modeling each aircraft emission. The number of operations was not specified for each individual aircraft. Therefore, the estimated concentrations downwind would be between the worst and the best case concentrations, if all model inputs are the same except source strength.

Response: No response needed.

<u>Reviewer 3</u>

Comment: Yes, the levels would decrease substantially. *Response:* No response needed.

8. Are there any other comments that you would like to make about the health consultation?

<u>Reviewer 1</u> *Comment:* No comment.

Reviewer 2 Comment: None.

<u>Reviewer 3</u> Comment: Yes. Please see attached pages. Response: ATSDR has responded to each comment in the following section.

9. Are there any comments on ATSDR's peer review process?

Reviewer 1 Comment: No comment.

Reviewer 2 Comment: None.

Reviewer 3 Comment: No.

10. Are there any other comments?

Reviewer 1

Comment: No comment.

Reviewer 2

Comment: Typo:

pp.72, please correct 'Figure B-8' to Figure B-9'.

pp. 40, 11th line of 4th paragraph, please delete 'and' in An individual's actual risk and may ...'

Response: These corrections were made.

Reviewer 3 Comment: No.

Reviewer 3 comments to question 8.

General Comments:

Comment: The main body of the text should include more detail on methodology, data, and maps. The reader finds considerable detail but often it is hidden in the footnotes of tables or is only included in the appendices. Conclusions seem to be valid, but a better presentation would lead to more reader confidence.

Response: ATSDR has adopted the present format on advice from health educators and risk communicators because the target audience of the health assessment is the general public, not the scientific community. Detail, whenever possible, is relegated to appendixes, footnotes, etc., to avoid distraction from the intended messages to the general public, but still included for the scientific readers.

Specific Comments:

Comment: Page 2, 1st Paragraph, Line 4: The sentence that begins "In an attempt" is slightly awkward and should be two sentences. *Response:* The sentence has been restructured.

Comment: Page 3, 1st Finding: The text does not seem to agree with Table 1. Perhaps it should state that analysis of hexavalent chromium before 1980 was not possible. As stated, it seems to imply that hexavalent chromium from stationary sources did have adverse health effects. Response: ATSDR has clarified the message.

Comment: Page 6, 1st Paragraph, Line 7: A conclusion of the study is presented in the background section with no support. Reasons should be given.

Response: The conclusion refers to current air emissions and the assessment and reasons for that conclusion are contained in that document.

Comment: Page 6, 1st Paragraph, Line 9: The reference to past air emissions (before 1995) is not supported. Earlier in the report (Table 1) past was also given as 1980. 1 think the text is

trying to tell me that only pre-1995 emissions were evaluated due to the change to JP-8, but I cannot be sure.

Response: The text states that pre-1995 air emissions were evaluated because of the use of JP-4 jet fuel before 1995. That point in time was necessarily used for all emissions even though chromium emissions changed in 1980. ATSDR initiated an investigation in 1996 and published a public health assessment (August 1999) addressing emissions from 1995 through base closure (2001).

Comment: Page 7, 2nd Paragraph, Line 1: I think the first sentence should be qualified by including "...at Kelly AFB." since this statement does not apply to the general literature. The next sentence stating distant past also has the same problem.

Response: This section applies to the general literature. For example, EPA's Air Toxics Monitoring Program began in the 1980s. ATSDR has modified the sentence to clarify.

Comment: Page 7, 4th Paragraph, Line 3: The sentence that begins on this line is confusing. It should be known if the chemicals were present or not. *Response:* It is not known at what level the chemicals were present.

Comment: Page 9, 1st Paragraph, Line 6: Why include such a long listing of chemicals if it is not complete?

Response: This listing identifies the chemicals for which data was provided.

Comment: Page 9, 1st Paragraph, Line 8: If most chemicals did not exceed health-based comparison values the important information is which ones did. This should be reworded or stated. Then the following sentence says no chemicals exceeded the noncancer comparison values while the next sentence said two did. A clean up is needed in this paragraph. **Response:** Health-based comparison values includes both cancer and noncancer comparison values. The text continues to specify that *noncancer* comparison values were *not* exceeded and two *cancer* comparison values were exceeded.

Comment: Page 9, 2nd Paragraph: This conclusion in bold font does not seem to agree with the previous paragraph.

Response: Exceeding a comparison value does not constitute a public health hazard, but identifies chemicals for which further evaluation is indicated.

Comment: Page 10, 1st Paragraph: The details on how much risk is involved and details should really be included in the main text and not only in the appendix. *Response:* See general comments response.

Comment: Page 10, 2nd Paragraph: Again the conclusion in **bold** font does not seem to agree with the previous paragraph.

Response: Exceeding a comparison value does not constitute a public health hazard, but identifies chemicals for which further evaluation is indicated.

Comment: Page 12, 1st Paragraph: What health outcomes were further evaluated and why? This is important information.

Response: The information is given in following paragraphs.

Comment: Page 12, 5th Paragraph: The maps should be presented in the main body of the text and would really help the reader.

Response: Because there may be multiple references to the same maps, ATSDR elects to locate maps in one place to avoid duplication.

Comment: Page 12, 6th Paragraph: If results are available, why not include them here? *Response:* Results are not yet available.

Comment: Page 13: Nice discussion, but what is needed is a summary paragraph of how these issues directly apply to this project.

Response: While the information may be relevant to this project, its direct relevance remains to be determined by followup activities and is best presented by those investigators. For example, the association of acute nonlymphocytic leukemia in children and parental occupational exposures to benzene may have relevance to the ZIP Code containing off-base housing and reporting elevated leukemia outcomes. Further investigation would be needed and the information is best presented in its entirety at that time. A summary statement has been added.

Comment: Page 17, 2nd Paragraph, Line 4: Site topography and geometry are also key inputs. *Response:* "Site topography and geometry" have been added.

Comment: Page 17, 2nd Set of Bullet Items, 1st Bullet: "....24 hours a day for any time period..." is a little confusing.

Response: The text was clarified by stating that the models can be used to estimate a substance's concentration for different time periods for which both emissions and meteorological data exist and that the ISCST model used in this report generates an hourly model result. The hourly results can then be compiled to generate maximum and average values over different time periods.

Comment: Page 18, 1st Paragraph: If the concentrations were measured, they did exist. I think what you may be trying to say is that these results are not applicable to all times or can be used for other nearby locations.

Response: The interpretation is as intended and no response is needed.

Comment: Page 18, Last Paragraph, Line 5: The sentence ending on this line should include "..... at the modeled locations." This is true because further downwind longer half-lives do make a difference.

Response: The text specifies that the point is near the base perimeter and that downwind concentrations will increase and thus "do make a difference." The text was changed to indicate that the fixed point off base was at the modeled location at the base perimeter.

Comment: Page 19, Figure A-1: Why does the last data point (~1500 minutes) does the concentration go down?

Response: The last data point was in error. The concentration was incorrectly entered as 0.00117 $\mu g/m^3$ and the correct value is 0.00122 $\mu g/m^3$. ATSDR also ran the model with half-lives of 280 minutes and 2,160 minutes to fill in the time between 60 and 500 minutes and to extend the end

of the curve. These additional points confirm the conclusion that the concentration becomes stable at the location of concern as the half-life approached 3 to 4 hours (180–240 minutes). Figure A-1 has been corrected and revised. The geographic location is in the Kelly Gardens community immediately north of and adjacent to Kelly AFB. The location in geographic coordinates is 641,600 meters west and 4,173,700 meters north in statewide grid, Texas South Central Zone, North American Datum of 1983.

Comment: Page 19, 1st Paragraph, Line 8: Should this be a factor be less than 50? 1 don't get this number from Figure A-2.

Response: The factor of 50 was the result of the change in the model parameters (release height, downwash, and rural versus urban dispersion) and the decrease in concentration as a function of distance within 300 meters of the source. Because this was not apparent, the number has been changed to 3 to represent the change in model parameters only and a factor of about 50 depending on the receptors location inside the base boundaries.

Comment: Page 19: This would also be the place to introduce which models were used, what inputs, were worst case or typical weather values used, etc. *Response:* ATSDR added this information.

Comment: Page 19, Caption for Figure A-2: Some parameters would have an effect. Those in the evaluation shown in the Figure perhaps have no effect.

Response: The caption for Figure A-2 has been changed to:

"Figure A-2. Input Parameter Comparison. Selection of model parameters shown in the Figure have no effect on off-base concentrations of contaminants, but may have significant effects upon on-base concentrations."

Comment: Page 21, Last Paragraph: Since hexavalent chromium is a solid, while the other pollutants are a gas, you may wish to mention modeling assumptions here, such as settling. *Response:* ATSDR added text to describe the assumption that all chemicals including metals were assumed to be in the gas form.

Comment: Page 22, 2nd Paragraph: More description of details would really help the reader. The levels that are used for comparisons (both criteria and measured) would indicate support for the statements made. The details do come out after a very careful review of Table B-1, but such important statements used in the paragraph should be supported in the text. *Response:* See response to general comment.

Comment: Page 24, 1st Paragraph: Just a thought, but could the efficiency of the scrubber used for control be included and possibly allow better interpretation of impacts before 1980? *Response:* Scrubber efficiency was considered but ATSDR determined that the uncertainty was too large to evaluate further. The uncertainty exists because the operation of the scrubbers in Building 301 changed in 1980 and four other plating shops existed prior to 1980.

The text states that scrubbers were installed in 1980 in Building 301. Additional information obtained by ATSDR shows that the scrubbers were installed when the building was constructed in 1977 with stack sampling tests in 1980. The text was clarified with this information (Kelly

AFB 2001). The scrubbers on Building 301 were originally designed to operate in a wet mode. However, a memorandum indicates that insufficient deionized water was available to operate the units so decisions were made to operate the units in a dry mode (Backlund 1995). The stack tests were completed in 1980 in a dry mode. The actual efficiencies prior to 1980 are not known. Emissions can be estimated from plating operations by knowing the level of plating operations but these data has not been identified.

Four other plating shops existed prior to 1980. The most significant one is the operation in Buildings 258/259 which began operation in 1942 and shutdown in 1977 (Kelly AFB 2001). The Air Force considered the information on past emissions from Buildings 258/259 incomplete and of low confidence because the buildings were demolished prior to this inquiry, there is missing data, lack of confidence in personnel interviews, and lack of construction drawings (EARTTECH 2000a, 2000b). As a result of these uncertainties, ATSDR did not evaluate impact before 1980.

References

Backlund 1995. Memorandum for Information from SA_ALC/EMC, Department of the Airforce, Headquarters San Antonio Air Logistics Center (AFMC), Kelly Air Force Base, Texas (R.J. Backland, P.E.) To SA-ALC/EMC (D.S. Guadarrama, P.E.), Subject Shutdown of Wet Scrubber Mode at Bldg 301 and Stack Sampling).

Kelly AFB 2001. Point Paper for Chromium Emission Data from Historical Plating Operations, Kelly AFB, Draft, June 26, 2001.

EARTHTECH 2000a. Final Report. Historical Air Emissions Estimate. Kelly AFB, TX. EARTH TECH, Inc. San Antonio, TX. March 27, 2000.

EARTHTECH 2000b. Addendum to the Historical Air Emissions Estimate Report, March 20, 2000. EARTH TECH, Inc. August 28, 2000. Transmitted by Charles Williams (Kelly AFB) on December 20, 2000

Comment: Page 25, 3rd Paragraph: The model used is included, but an earlier introduction would be better for reader understanding. Also, how the model was used should be mentioned. Details are given later but the fact a volume approach using ISCST with estimated positions rather than using EDMS should be discussed.

Response: The following changes, shown in italics with accompanying text were made to address this comment.

Page 8 – First Paragraph

Data on JP-4 jet fuel speciation acquired by ATSDR and operational data provided by Kelly AFB were used to conduct an air dispersion model of aircraft emissions. A worst-case jet fuel emissions scenario was used for modeling aircraft emissions. *The Industrial Source Complex air dispersion model was used (ISCST3, see Appendix B for details)*.

Page 25. Model Inputs The Industrial Source Complex-Short Term (ISCST) model was used to perform the air modeling. To use this model, information on the source of pollutants, ambient meteorology, and information on receptor locations must be entered into the model. The model simulates the movement of the pollutants in the atmosphere and calculates a concentration at the given receptor locations. *The emissions were treated as a series of volume sources behind the aircraft (see page 32 for details)*.

Bottom of Page 33 and Top of Page 34.

Forty-eight volume sources were used to represent taxiway emissions. Fourteen were used to represent takeoffs. Thirty were used to represent climbout. Eighty were used to represent approach. These sources represent aircraft movement at approximately 3-second intervals. Sources in each category were spaced according to their respective speed during that mode.

The commenter suggests that the EDMS model should be discussed in conjunction with the use of the ISCST model with volume sources. The ISCST model was used for modeling jet aircraft emissions using volume sources. The size and location of these volume sources were estimated and discussed in Appendix B. The use of the EDMS model at the time the modeling for this report was developed (March 2001 through June 2001) was not an option. In March 2001, version 3 was the current version of the EDMS model. EDMS version 4 was released in May 2001. The EDMS models (Versions 3 and 4) were developed for criteria air pollutants plus hydrocarbons (carbon monoxide, nitrogen oxides, sulphur oxides, hydrocarbons, and suspended particles). The EDMS V3 was not easily adaptable for other emission factors and chemicals including hazardous air pollutants that were the subject of this study. EDMS currently was not designed to perform air toxic analyses for aviation sources, but could have been supplemented with other air toxic methodology and models ([Federal Register: October 13, 1999 (Volume 64, Number 197)] [Notices] [Page 55525–55595]). The EDMS V4 model now has the flexibility to import the emission factors for new aircraft and additional chemicals. The EDMS V3 did not.

EDMS V3 used two models called PAL2 and CALINE3 that simulated aircraft emissions as line sources. CALINE3 is a line source model and assumes that a zone (volume) containing the line source exists with the zone. The size of the zone is a function of line width and an initial vertical dispersion. The contaminants in this zone then undergo vertical and horizontal dispersion using a steady-state Gaussian model (Benson 1979). PAL2 calculates a line source by integration of point sources (EPA 1978). The location accuracy of the points and lines representing the planes and the relative accuracy when compared to ATSDR's volumes sources is not known.

EDMS V4 is a significant revision of EDMS V3. EDMS V4 uses EPA's AERMOD air dispersion model. AERMOD in EDMS uses areas sources for aircraft taxiing, aircraft queuing, aircraft accelerating on the runway, aircraft after takeoff and during the landing approach. The area source was selected, as opposed to using a series of volume sources based on recommendations from the American Meteorological Society/EPA Regulatory Model Improvement Committee (AERMIC) (CSSI Inc. 2002). A comparison of the EDMS V4 model with ATSDR's approach is possible because the U.S. Army Center for Health Promotion and Preventive Medicine (CHPPM), in the fall of 2003, used the EDMS V4 to evaluate ATSDR's results. Using ATSDR's assumption of 336,000 operations of the B52H and the emission rates identified in ATSDR's report, CHPPM's result at the point of maximum impact were within 10% of ATSDR's result (personal communication, Les Pilcher, US Army Center for Health Promotion and Preventive Medicine, December 19, 2003, [49]. This indicates that the different models using the same assumptions have good agreement.

References

Benson, Paul 1979. Abridged version of "CALINE3 - A Versatile Dispersion Model for Predicting Air Pollutant Levels Near Highways and Arterial Streets-Interim Report" Report Number FHWA/CA/TL-79/23, Nov. 1979. Paul E. Benson, Office of Transportation Laboratory, California Department of Transportation. Abridged Version by Computer Sciences Corporation [http://www.epa.gov/scram001/userg/regmod/caline3.pdf].

CSSI, Inc. 2002. Emissions and Dispersion Modeling System (EDMS) Reference Manual. Prepared for U.S. Department of Transportation, Federal Aviation Administration, Washington, D.C. Document Number FAA-AEE-01-01

US Environmental Protection Agency (EPA). 1978. User's guide for PAL, A guassian-plume algorithm for point, area, and line sources. Research Triangle Park (NC): Office of Research and Development, Environmental Sciences Research Laboratory, February. Report No. EPA-600/4-78-013.

Comment: Page 26, 5th Paragraph: Listing of all these aircraft would seem to be better in a table. *Response:* The list was revised as a table.

Comment: Page 26, 5th Paragraph: If only the B52H and F16 are to be used, why include that emission data exist for all of these other aircraft were found?

Response: The list of aircraft with available emissions data was provided to demonstrate the process ATSDR went through to reconstruct past exposures. The list more clearly demonstrates the advantages and disadvantages of the assumption to only use the B52-H and the F16. It also provides readers with knowledge of the data that is available. No change was made in the text.

Comment: Page 27, Table B-2: The numbers look too high. Could a B52H emit 113 kilograms of hydrocarbons during the taxi-out? This is true of Table B-3 as well. Also, if Touch & Go are not going to be used, why include?

Response: ATSDR checked the source document and it indeed lists 113 kilograms of hydrocarbons (HCs) during the taxi-out (Seitchek 1985). This number does seem high. It was derived based on time-in-mode, engine setting, and HC emission rate. The power setting for taxi-out is idle which has the highest HC emissions rate. For the TF33-3 engine, the rate is 84 g HC/kg fuel. The fuel rate is 0.11 kg/s so the HC emission rate is 9.24 g/s. For 113 kilograms, the time-in-mode for taxi-out would need to be 3.4 hours which seems very unrealistic. ATSDR checked the KC-135A from this reference for taxi-out and came up with 11.5 hours which is even more unrealistic. ATSDR suspects a systematic error in Table A of Seitchek (1985). It's possible that the units for the table are kilograms and not metric tons. A note was added to Table B-2 about this possible error.

Because the values in Table B-2 were only used for a comparison among planes and not used in the emissions modeling, the error in Seitchek (1985) does not change our results. The hydrocarbon rates used in the modeling was 94 g HC/kg fuel and 0.14 kg/s of fuel (Spicer et al 1988). These values are similar to Seitchek (1985). The times-in-mode used in the ATSDR modeling was 9 minutes for taxi-out (Naugle et al 1975) for a total of 7.1 kg HC released during taxi-out.

References

Aircraft Emissions Characterization." C.W. Spicer, M.W. Holdren, S.E. Miller, D.L. Smith, R.N. Smith, D.P. Hughes. Final Report, March 1988, Engineering and Services Laboratory, Air Force Engineering & Services Center, Tyndall Air Force Base, ESL-TR-87-63.

USAF Aircraft Engine Emissions Estimator, Glenn D. Seitchek, ESL-TR-85-14, November 1985.

USAF Aircraft Pollution Emission Factors and Landing and Takeoff (LTO), Dennis Naugle, et al, AD/A-006 239 (February 1975)

Comment: Page 29, Equation at Top of Page: Derivation of this formula should be discussed. I get different answers when using moles for mass conversion.

Response: The equation is

The equation is

(1)
$$\% wt \frac{HAP}{HC} = \left(\frac{[HAP]}{[HC]}\right) \times \left(\frac{NumberofC_{HC}}{NumberofC_{HAP}}\right) \times \left(\frac{MW_{HAP}}{MW_{HC}}\right)$$

where:

 $[HAP] = \text{concentration of organic compound in ppm_vC}$ $[HC] = \text{concentration of total hydrocarbons in ppm_vC}$ NumberofC_{HC} = Number of carbon molecules, 9.3 is used for HC* NumberofC_{HAP} = Number of carbon molecules in the HAP $MW_{HAP} = \text{Molecular Weight of the HAP}]$ $MW_{HC} = \text{Molecular weight of the total hydrocarbons} = 130*$ *Douglas, Everett, Naval Aviation Depot, Naval Air Station, San Diego, California. Email record of personal communication regarding information about converting units and data on the number of carbons and molecular weights of total hydrocarbons in jet fuel, February 12, 2001.

The units of concentration in this formula require ppm_vC . Moles should not to be used in this formula as it is based on a volume per volume basis. The formula was taken from AESO (1999) and the text will be referenced accordingly. There was a typographically error and OC was changed to HAP and the subscript "v" was added to indicate that it is based on volume not mass.

The derivation of the formula is based on two equations:

(2) $ppm_v = \frac{ppm_v}{\#C}$
where $ppm_v = parts per million by volume$ <math>ppmC = parts per million carbon#C = number of carbons in the molecule

and the ideal gas law,

PV = nRT,

which is used to convert ppm, to a mass basis.

Where P = pressure of the gas V = the volume it occupies T = its temperature (in Kelvin) n = number of moles of gas present R = universal gas constant

First, using the ideal gas law, the volume of 1 mole of air (V) is calculated.

(4)
$$V = R\left(\frac{L-atm}{{}^{\circ}K-Mole}\right) \times \frac{T({}^{\circ}K)}{1atm} = \frac{L}{mole}$$

At standard temperature and pressure (273°K and 1 atm). The volume is

(5)
$$V_{STP} = 0.8206 \left(\frac{L - atm}{{}^{o}K - Mole} \right) \times \frac{273^{o}K}{1atm} = 22.4 \frac{L}{mole}$$

ppm_v is defined as

(6)
$$ppm_v = \frac{\mu L_{HAP}}{L_{air}}$$

where V_{air} is the total volume of air, V_{HAP} is the volume of air occupied by the HAP Using dimensional analysis and equations 5 and 6.

(7)
$$ppm_{\nu} = \frac{\mu L_{HAP}}{L_{air}} \times \frac{10^3}{1m^3_{air}} \times \frac{1L_{HAP}}{10^6 \mu L_{HAP}} \times \frac{mole}{22.4 L_{HAP}} \times \frac{MW}{mole} \times \frac{g}{mole} \times \frac{10^3 mg}{g}$$

or

(8)
$$ppm_{v} = \left(\frac{MW_{HAP}}{22.4L_{HAP}}\right) \frac{mg}{m^{3}}$$

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Combining with equation 1 and solving for mass

(9) concentration in mg/m³ =
$$\frac{ppm_{\nu}C_{HAP}}{\#C_{HAP}} \times \left(\frac{MW_{HAP}}{22.4L_{HAP}}\right) \frac{mg}{m_{air}^3}$$

To obtain a weight ratio of HAP to hydrocarbon (HC)

(10)
$$\%wt \left(\frac{HAP}{HC}\right) = \frac{\frac{ppm_{v}}{\#C_{HAP}} \times \left(\frac{MW_{HAP}}{22.4L_{HAP}}\right) \frac{mg}{m^{3}_{air}}}{\frac{ppm_{v}C_{HC}}{\#C_{HC}} \times \left(\frac{MW_{HC}}{22.4L_{hC}}\right) \frac{mg}{m^{3}_{air}}}$$

Simplifying and rearranging brings us back to equation 1.

(11)
$$\%wt\left(\frac{HAP}{HC}\right) = \frac{\frac{ppm_{v}C_{HAP}}{\#C_{HAP}} \times (MW_{HAP})}{\frac{ppm_{v}C_{HC}}{\#C_{HC}} \times (MW_{HC})} = \frac{ppm_{v}C_{HAP} \times (MW_{HAP}) \times \#C_{HC}}{ppm_{v}C_{HC} \times (MW_{HC}) \times \#C_{HAP}}$$

Note that this equation is independent of temperature and pressure.

References

Aircraft Environmental Support Office (AESO), 1999. Toxic Organic Contaminants in the Exhaust of Gas Turbine Engines for JP-5 and JP-8 Fuel: Draft. San Diego (CA): U.S. Navy, Aircraft Environmental Support Office Naval Support Depot-North Island. AESO Report No. 12-90, Revision B. February.

Comment: Page 30, Continuation of Table: Headings should be included at the top of the page. *Response:* The heading was added.

Comment: Page 33, 2nd Paragraph: Do you mean Touch & Go operations were divided equally among takeoffs and landings?

Response: Touch and go operations were not specified in the number of annual operations. Since the annual operations most likely did include touch and go operations, ATSDR took the most conservative approach (highest emissions) and assumed that the unknown number of touch and go operations was a takeoff or a landing but not both. This means that the 336,000 operations were divided into 168,000 takeoffs and 168,000 landings. This text was added to the report for clarity.

Comment: Page 33, 9th Paragraph: "..... from about 480 meters." Does this mean to the ground, around this height, or something else?

Response: This meant that source release heights for approach varied from about 480 to 0 meters above ground. The text was modified for clarification.

Comment: Page 33, 11th Paragraph-. 46 minutes seems like a very long taxi time.

Response: The taxi time is the total time for taxi during takeoff and taxi during landing (see Table B-4)and includes time for startup (20 minutes), outbound taxi (9 minutes), inbound taxi (12 minutes), and idle at shutdown (4.8 minutes). This data was obtained from USAF Aircraft Pollution Emission Factors and Landing and Takeoff (LTO), Dennis Naugle, et al, AD/A-006 239 (February 1975). The text was clarified accordingly.

References

USAF Aircraft Pollution Emission Factors and Landing and Takeoff (LTO), Dennis Naugle, et al, AD/A-006 239 (February 1975)

Comment: Page 34: I expected information on the F- 16 to also be presented instead of just the B52H.

Response: This page does include information on the F16.

Comment: Page 34, 2nd Paragraph: It would be good to tell the reader what the level where health effects begin is and from what reference in this paragraph. *Response:* See response to general comment and Figures B-6 and B-7.

Comment: Page 44, Figure B-7: The public may have the wrong idea of what is toxic when outdoor levels at Kelly AFB are reported to be above those in a smoked-fill bar. Some description in the text may help.

Response: ATSDR objectively presents information reported in the scientific literature to give a complete perspective.

Comment: Page 46: It is confusing to have two Appendix Bs. Perhaps B1 and B2? *Response:* There is only one Appendix B, containing two attachments.

Comment: Page 55, 1st Paragraph, Line 3: The words "...near Kelly AFB." after the word "...susceptibility...." may be called for. *Response:* This is a general statement not specific to Kelly AFB.

Comment: Page 55, 5th Paragraph: This paragraph should be included in main body of text. *Response:* ATSDR prefers that these general methodological statements remain with other like statements in an appendix than inserted into a discussion of findings in the main body of text.

Comment: Page 56, 1st Paragraph: How risk factors were developed is not completely described. More detail would be helpful. *Response:* More detail was added to the tables and text.

Comment: Page 59, 4th Bullet Item: Is there a low to moderate increase in risk or a low to moderate risk?

Response: There is a low to moderate increase in estimated risk over the background risk.

Comment: Page 60, Table C- 1: That 6 people in 1000 would be at cancer risk is quite high. This needs more discussion in the text as does even 3 people in 1000. *Response:* Clarification has been added to text.

Comment: Page 64, Reference 40: A much more recent reference exists. *Response:* ATSDR agrees that a more recent reference exists; however, the data came from Reference 40 because it was the document that was available at the time.

Comment: Page 66, 4th Paragraph: It would be good to report measured values. **Response:** The text in this paragraph referred to concentrations reported from the air toxics monitor located about 10 to 15 miles northeast of Kelly AFB at 254 Seale Road, San Antonio, Texas. The concentrations are shown below. These values are annual maximum numbers. Detection levels were used when a compound was not detected. This table and the information about the monitor was not included in the revised report (response to peer review comments) because this data was provided for clarification purposes only and do not impact the results and conclusions on past air emissions.

Year	Butadiene		Benzene		
	mg/m	Risk	mg/ m ³	Risk	Total Risk
1994	0.17	4.73E-05	2.15	0	0.0000641
1995	0.71*	1.98E-04	1.63	1.27E-05	2.11E-04
1996	0.71*	1.98E-04	1.38	1.08E-05	2.09E-04
1997	0.71*	1.98E-04	1.76	1.37E-05	2.12E-04
1998	0.71*	1.98E-04	1.46	1.14E-05	2.09E-04
1999	0.74	2.07E-04	1.42	0	2.18E-04
2000	0.10	2.93E-05	1.09	0	3.78E-05
2001	0.11	3.15E-05	1.57	1.22E-05	0.0000437
Average	0.49	0	1.56	0	0.000151

Inhalation unit risk used for butadiene = $0.00028 \text{ (mg/m}^3)^{-1}$ Inhalation unit risk used for benzene = $0.0000078 \text{ (mg/m}^3)^{-1}$

* Detection level used

Comment: Page 67, 4th Paragraph: What was the logic for only using a 9 hour half-life? This doesn't appear to follow from the table above the paragraph.

Response: ATSDR concurs that the table and text are not clear. 1,3-butadiene is estimated to have a short atmospheric lifetime because of its reactivity. The actual lifetime depends upon the conditions at the time of release. The primary removal mechanisms are through chemical reactions with hydroxyl radicals and ozone. Therefore, factors influencing 1,3-butadiene's atmospheric lifetime, such as the time of day, sunlight intensity, temperature, etc., also include those affecting the availability of hydroxyl radicals and ozone (EPA 1993). The literature reports different half-lives and in many cases, do not specify controlling factors that would influence butadiene degradation. The Table was an attempt to show the half-lives as a function of a single factor.

For the modeling, ATSDR used a report by the California Air Resources Board that stated "[a]tmospheric half-lives of 1 to 9 hours are expected." (CARB 1997). This range was

reasonable to evaluate as 1 hour was near the lower end reported. Nine hours was reasonable to use as a higher value because it is in the range of the higher values. Higher half-lives would not significantly change the concentrations near the base where the population of interest resides because the travel of air emissions time is much faster than 9 hours (Figure A-1 demonstrated this for hexavalent chromium). The model was run with no degradation as a worst case.

ATSDR clarified this in the text and merged this discussion with text in Appendix B under a new section called *Sensitivity Analysis*.

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Comment: Page 69, Table B-8: Again, high risk values are reported and probably need to be discussed more.

Response: Clarification has been added to text.

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Kelly Special Restoration Advisory Board (RAB) Meeting Agenda Tuesday, November 09, 2004 6:30 p.m. Kennedy High School Cafeteria 1922 S. General McMullen Drive

Meeting Goals

The RAB will:

- Review and comment on the ATSDR Health Consultation, Past Air Emissions Report for Kelly Air Force Base
- I. <u>Roll call begins at 6:30 p.m.</u>
 - Meeting will convene
 - Pledge of Allegiance
 - Moment of silence
 - Welcome and introductions
 - Discuss goals for this meeting
- II. A. Agency for Toxic Substances and Disease Registry (ATSDR) Community Update/ Past Air Emissions Report
 - B. Community Comment/Question-and-Answer session on the ATSDR Community Update/ Past Air Emissions Report
- III. A. TAPP Update (if time allows)
 - B. Community Comment/Question-and-Answer session on the TAPP update
- IV. Meeting Wrap-Up
 - Action items will be addressed at the next regular RAB meeting
 - Meeting summaries will be approved at the next regular RAB meeting
 - Next TRS meeting proposed for Tuesday, December 13, 2004, at 6:30 p.m., at the Environmental Health and Wellness Center*
 - Next RAB meeting proposed for Tuesday, January 18, 2004, at 6:30 p.m., at the Kennedy High School Auditorium *

*Meeting dates and locations are subject to change.

Note:

• Copies of the ATSDR Past Air Emissions Report are available at this meeting. The public comment period extends through Nov. 30, 2004. Comments on this report must be made in writing. Mail comments to: Chief, Information Services Branch

ATSDR 1600 Clifton Road, N.E. (MS E-60)

- Atlanta, GA 30333
- RAB elections are January 18, 2004. The application deadline is January 4, 2004. Pick up an application at this meeting or call 925-0956 to request an application.

Dr. David Smith

Ms. Larisa Dawkins

Dr. David Smith

Ms. Maria Teran-McIver Dr. David Fowler

Dr. David Smith

Dr. David Smith

Kelly Special Restoration Advisory Board (RAB) Technical Review Subcommittee (TRS) SUMMARY

November 9, 2004 Kennedy High School Cafeteria 1922 S. General McMullen Dr. San Antonio, TX 78226

1. Attendees:

Mr. Don Barker Ms. Sandra Converse Ms. Kyle Cunningham Ms. Larisa Dawkins Ms. Leigh-Ann Fabianke Dr. David A. Fowler Mr. Ramon A. Garcia Mr. Rodrigo Garcia Ms. Sarah Garcia Mr. Henry Galindo Ms. Esmeralda Galvan Ms. LeAnn Herren Mr. Brian Howard Ms. Jill Johnston Mr. Brian M. Kaplan Ms. Cheri Kirkpatrick Ms. Norma Landez Ms. Maria Teran-McIver Mr. Carlos San Miguel Mr. Gary Miller

Mr. Sam Murrah Mr. Pete Muzquiz Ms. Jannie O'Neal Mr. Nazarite Perez Ms. Abbi Power Mr. Armando Ouintanilla Mr. Genaro Rendon Mr. Sam Sanchez Mr. Jeffrey Shire Mr. Michael Sheneman Mr. Robert Silvas Mr. Kelley Siwecki Mr. Brendan Smith Dr. David Smith Mr. Tim Sueltenfuss Ms. Robyn Thompson Mr. Glenn Wilkenson Mr. James Wittmer Mr. David Yantz

- 2. Introduction. Dr. David Smith, RAB/TRS Facilitator, opened the meeting at 6:35 p.m. and called the roll. A quorum was achieved just after the meeting began. Dr. David Smith announced that Mr. William Ryan would be filling Mr. Adam Antwine's seat as government co-chair for this Special RAB meeting. The Pledge of Allegiance was said and then a moment of silence was observed. Dr. David Smith announced that the goal of the meeting was to review and comment on the Agency for Toxic Substances and Disease Registry (ATSDR) Health Consultation, Past Air Emissions Report for Kelly Air Force Base.
- 3. ATSDR Briefing. Dr. David A. Fowler and Mr. Brian M. Kaplan of the ATSDR presented a briefing on the ATSDR Health Consultation, Past Air Emissions Report for Kelly Air Force Base.
- 4. Community Comment/Question-and-Answer Session on the ATSDR Past Air Emissions Report. Dr. David A. Fowler, Mr. Brian M. Kaplan and Ms. Maria Teran-McIver responded

to questions from the audience and RAB members in attendance concerning the ATSDR Report and related subjects.

- 5. Technical Assistance for Public Participation (TAPP) Update. Ms. Larisa Dawkins presented a briefing on what the TAPP is and how it works. She also provided a status update of the TAPP program budget.
- 6. Community Comment/Question-and-Answer Session on the TAPP update. Ms. Larisa Dawkins responded to questions from the audience and RAB members in attendance concerning the TAPP Update and related subjects.
- 7. Meeting Wrap-Up. Dr. David Smith stated that action items will be discussed at the next regularly scheduled RAB meeting. Meeting summaries will also be approved at the next regular RAB meeting. Dr. David Smith also reminded the RAB members and the community that RAB elections will take place at the January 18, 2005 RAB meeting.
- 8. Next Meeting. The next regularly scheduled RAB meeting is set for Tuesday, January 18, 2005, at 6:30 p.m. at Kennedy High School. The next regularly scheduled TRS meeting is set for Monday, December 13, 2004, at 6:30 p.m. at the Environmental Health and Wellness Center.
- 9. Adjourn. The meeting adjourned at 9:45 p.m.

Consejo Consultivo para la Restauración Especial de Kelly (RAB, por sus siglas en inglés) Subcomité de Revisión Técnica (TRS, por sus siglas en inglés) RESUMEN

9 de noviembre de 2004 Cafetería de la Preparatoria Kennedy 1922 S. General McMullen Dr. San Antonio, TX 78226

1. Asistentes:

Sr. Don Barker Sra. Sandra Converse Sra. Kyle Cunningham Sra. Larisa Dawkins Sra. Leigh-Ann Fabianke Dr. David A. Fowler Sr. Ramon A. Garcia Sr. Rodrigo Garcia Sra. Sarah Garcia Sr. Henry Galindo Sra. Esmeralda Galvan Sra. LeAnn Herren Sr. Brian Howard Sra. Jill Johnston Sr. Brian M. Kaplan Sra. Cheri Kirkpatrick Sra. Norma Landez Sra. Maria Teran-McIver Sr. Carlos San Miguel Sr. Gary Miller

Sr. Sam Murrah Sr. Pete Muzquiz Sra. Jannie O'Neal Sr. Nazarite Perez Sra. Abbi Power Sr. Armando Quintanilla Sr. Genaro Rendon Sr. Sam Sanchez Sr. Jeffrey Shire Sr. Michael Sheneman Sr. Robert Silvas Sr. Kelley Siwecki Sr. Brendan Smith Dr. David Smith Sr. Tim Sueltenfuss Sra. Robyn Thompson Sr. Glenn Wilkenson Sr. James Wittmer Sr. David Yantz

- 2. Introducción. Dr. David Smith, Intermediario de RAB/TRS, hizo la apertura de la junta a las 6:35 p.m. y pasó lista. Se alcanzó el quórum exactamente después de que empezó la junta. El Dr.David Smith anunció que el Sr. William Ryan ocuparía el puesto del Sr.Adam Antwine como co-dirigente del gobierno para esta junta Especial de RAB. Se hizo el Juramento de Lealtad y después se guardó un momento de silencio. El Dr. David Smith anunció que la meta de la junta era revisar y comentar sobre la Consulta de Salud de la Agencia para Sustancias Tóxicas y Registro de Enfermedades (ATSDR, por sus siglas en inglés), el Reporte Pasado de Emisiones de Aire de la Base Kelly de la Fuerza Aérea.
- **3.** Informe del ATSDR. El Dr. David A. Fowler y el Sr. Brian M. Kaplan de ATSDR presentaron un informe sobre la Consulta de Salud de ATSDR, el Reporte Pasado de Emisiones de Aire de la Base Kelly de la Fuerza Aérea.
- 4. Comentarios de la Comunidad / Sesión de Preguntas y Respuestas sobre el Reporte Pasado de Emisiones de Aire de ATSDR. El Dr. David A. Fowler, el Sr. Brian M. Kaplan y

la Sra. Maria Teran-McIver respondieron las preguntas de la audiencia y de los miembros de RAB que asistieron, concernientes al Reporte de ATSDR y asuntos relacionados.

- 5. Actualización de la Asistencia Técnica para la Participación Pública (TAPP, por sus siglas en inglés.) La Sra. Larisa Dawkins presentó un informe sobre lo que es TAPP y cómo funciona. También proporcionó una actualización de la situación del presupuesto del programa TAPP.
- 6. Comentarios de la Comunidad / Sesión de Preguntas y Respuestas sobre la actualización de TAPP. La Sra. Larisa Dawkins respondió las preguntas de la audiencia y de los miembros de RAB que asistieron, concernientes a la Actualización de TAPP y asuntos relacionados.
- 7. Terminación de la Junta. El Dr. David Smith estableció que las acciones a tomar, serán discutidas en la siguiente junta de RAB programada regularmente. Los resúmenes de la junta también serán aprobados en la siguiente junta regular de RAB. El Dr. David Smith también les recordó a los miembros de RAB y a la comunidad, que las elecciones de RAB se llevarán a cabo en la junta de RAB del 18 de enero de 2005.
- 8. Próxima Junta. La siguiente junta de RAB programada regularmente, será el martes 18 de enero de 2005, a las 6:30 p.m. en la Preparatoria Kennedy. La siguiente junta de TRS programada regularmente será el lunes 13 de diciembre de 2004, a las 6:30 p.m. en el Centro de Higiene y Bienestar Ambiental.
- 9. Cierre. Se levantó la sesión a las 9:45 p.m.

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The Agency for Toxic Substances and Disease Registry (ATDSDR) invites you to an Open House about Kelly AFB

> December 14, 2004 4:00 p.m. - 8:00 p.m. Kennedy High School Cafeteria 1922 S. General McMullen Dr. San Antonio, TX 78226

ATSDR invites you to drop by anytime between 4 p.m. and 8 p.m. to discuss ATSDR activities concerning contamination issues related to Kelly AFB.

ATSDR staff will be available to speak with you one-on-one about the agency's activities and about public health activities conducted in the Kelly AFB area.

ATSDR has completed evaluations of: soil, water, air and fish

ATSDR is finalizing: Evaluation of East Kelly, Evaluation of health outcome data, Past Air Emissions at Kelly AFB

Staff from San Antonio Metro Health Dept. and Texas State Health Services Dept. also will be available to talk with you about their involvement with Kelly AFB.

Handouts will be available in Spanish

La Agencia para Sustancias Tóxicas y Registro de Enfermedades, ATSDR, les invita a una Sesión de Puertas Abiertas

14 de diciembre 4:00 p.m. - 8:00 p.m. en la Cafeteria de la Kennedy High School 1922 General McMullen Dr. San Antonio, TX 78226

El público puede llegar en cualquier momento entre los 4 p.m. y 8 p.m.para discutir las actividades de la ATSDR en caso de la contaminación relacionada con la Base de Kelly AFB.

Personal de la ATSDR estará disponible para hablar uno en una con miembros de la comunidad sobre las actividades de la ATSDR.

Las investigaciones terminadas incluyen: La evaluación del suelo, del agua, del aire y de los pescados.

Las actividades todavía en plazo incluyen: Evaluación de East Kelly, Evaluación de Datos de Salud, y Las Emisiones de Aire en el Pasado.

Miembros del personal del departamento de la salud del metro de San Antonio y el departamento de los servicios médicos del estado de Tejas también estarán disponible para el público para hablar sobre su implicación en la obra de evaluación de salud.

Hojas informativas serán disponibles en español. The Agency for Toxic Substances and Disease Registry (ATDSDR) invites you to an Open House about Kelly AFB

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DEPARTMENT OF THE AIR FORCE AIR FORCE REAL PROPERTY AGENCY

6 December 2004

Ms. Sonja S. Coderre Public Affairs Officer 143 Billy Mitchell Blvd Ste 1 San Antonio TX 78226-1816

Dear Restoration Advisory Board (RAB) Members

Thank you for your continued interest in the Kelly environmental cleanup program. For your reference, I have included a summary of the 9 November 2004 Special RAB meeting.

This summary is a brief overview of what occurred at the Special RAB meeting. A court reporter prepared a word-by-word transcript of the RAB meeting that will be made available for review. You will receive a copy of that transcript at the 18 January RAB meeting. If you would like to request a copy of the transcript prior to the January meeting, please call 925-0956. After the transcript has been approved, a copy will be placed at the following Information Repositories:

- San Antonio Central Public Library
 600 North Soledad, 2nd Floor Government Documents Section San Antonio TX 78205
- Former Kelly AFB Library
 250 Goodrich Drive, Bldg 1650, Room 138
 San Antonio TX 78226
- Environmental Health and Wellness Center 911 Castroville Road San Antonio TX 78237

I appreciate the opportunity to share information on the Kelly environmental cleanup program with you. If you have any questions, please feel free to contact us at (210) 925-0956.

Sincerely

Attachment: 9 November 2004 Special RAB Meeting Summary

























ADMINISTRATIVE RECORD

FINAL PAGE